

## The Pacific Forestry Centre, Victoria, British Columbia

The Pacific Forestry Centre of the Canadian Forest Service undertakes research as part of a national network system responding to the needs of various forest resource managers. The results of this research are distributed in the form of scientific and technical reports and other publications.

Additional information on Natural Resources Canada, the Canadian Forest Service, and Pacific Forestry Centre research and publications is also available on the World Wide Web at: **www.pfc.cfs.nrcan.gc.ca.** To download or order additional copies of this publication, see our online bookstore at: **bookstore.cfs.nrcan.gc.ca.** 

# Modeling the effect of landscape pattern on mountain pine beetles

Josie Hughes<sup>1,\*</sup>, Andrew Fall<sup>1,2</sup>, Les Safranyik<sup>3</sup> and Ken Lertzman<sup>1</sup>

<sup>1</sup>School of Resource and Environmental Management Simon Fraser University 8888 University Drive Burnaby, BC, V5A 1S6, Canada

<sup>2</sup>Gowlland Technologies Ltd Victoria, BC

<sup>3</sup> Natural Resources Canada Canadian Forest Service Pacific Forestry Centre Victoria, BC

\* Present address: Department of Zoology, University of Toronto 25 Habord St., Toronto, ON, M5S 3G5 (416-534-0827. js.hughes@utoronto.ca)

Natural Resources Canada Canadian Forest Service Pacific Forestry Centre Information Report BC-X-407 Canadian Forest Service Pacific Forestry Centre 506 West Burnside Road Victoria, British Columbia V8Z 1M5 Phone (250) 363-0600

www.pfc.cfs.nrcan.gc.ca

© Her Majesty the Queen in Right of Canada, 2006

ISSN 0830-0453 ISBN 0-662-44440-X Printed in Canada

Microfiches of this publication may be purchased from: MicroMedia Inc. 240 Catherine Street, Suite 305 Ottawa, ON K2P 2G8

Library and Archives Canada Cataloguing in Publication

Modeling the effect of landscape pattern on mountain pine beetles / Josie Hughes ... [et al.].

(Information report; BC-X-407) Includes bibliographical references: p. Includes abstract in French. ISBN 0-662-44440-X

Cat. no.: Fo143-2/407E

1. Mountain pine beetle--Effect of forest management on. 2. Mountain pine beetle--Control--Simulation methods. 3. Mountain pine beetle--Behavior. 4. Trees--Diseases and pests--Control. I. Hughes, Josie, 1978- II. Pacific Forestry Centre III. Series: Information report (Pacific Forestry Centre) BC-X-407.

SD945.M78M72 2007

634.9'7516768

C2006-980281-5

# **Contents**

ABSTRACT / KESUME	V
1.0 Introduction	1
1.1 Management context and motivation	
1.2 Terminology	1
1.3 Expectations	1
1.3.1 EFFECT OF PATTERN IN GENERAL	1
1.3.2 Consequences of Pheromone mediated aggregation	3
1.3.3 Consequences of free flight	4
1.4 Overview and objectives	4
2.0 Methods	6
2.1 The model	6
2.2 Experimental design and analysis	6
2.2.1 PATTERN GENERATION AND EXPERIMENTAL DESIGN	6
2.2.2 Analysis Procedure	7
3. Experiments and results	10
3.1. The model	10
3.1.1. Notation	10
3.1.2. Conceptual overview of the flight model	10
3.1.3. Mountain pine beetle mode changes	10
3.1.4. Mountain pine beetle movement	12
3.1.5. LANDING AND PHEROMONE PRODUCTION	13
3.1.6. Pheromone diffusion and decay	14
3.1.7. Complete model of dispersal, aggregation and attack	15
3.1.8 Extending the model over multiple years	15
3.1.9. Base model parameterization	16
3.1.10 Numerical methods and model implementation	19
3.1.11 Temporal extent and resolution	20
3.1.12 Initial conditions	20
3.2 Clarifying hypotheses	
3.3 Experiment I – base case	
3.4 Experiment II – the effect of free flight	
3.5 Experiment II extension – sensitivity without free flight	
4. Discussion	
4.1. Beetle ecology	
4.2. Movement modeling	
ACKNOWLEDGEMENTS	
References	
Tables	
Appendix	44

# **Tables**

1a.	Definitions of terms	38
1 <sub>B</sub> .	Definitions of terms (continued).	39
2.	PROBABILITIES FOR TRANSITIONS BETWEEN FIVE BEETLE MODES	40
<b>3.</b>	MODEL PARAMETERS AND THEIR ASSOCIATED BASE VALUES.	41
4.	SUMMARY OF EXPERIMENTAL RESULTS	43
	Figures	
1.	THE POTENTIAL EFFECTS OF CHANGING HABITAT CONFIGURATION	2
2.	EXPERIMENTAL DESIGN SCHEME.	7
3.	SENSITIVITY OF WEIBULL-TYPE CURVE TO SHAPE PARAMETERS	8
4.	SUMMARY OF MODEL WITHIN THE FLIGHT PERIOD	11
<b>5.</b>	DISTANCES FLOWN BY MODELED MOUNTAIN PINE BEETLES	18
6.	SENSITIVITY OF PHEROMONE PLUME SIZE AND SHAPE	19
7.	DEFINITION AND CALCULATION OF CENTER CLUSTER SIZE.	21
8.	THE EFFECT OF HABITAT COMPACTION, PATCH SIZE, AND COMMUNICATION DISTANCE	
	ON CENTER CLUSTER SIZE	22
9.	PHEROMONE PLUMES USED IN BASE EXPERIMENT	23
10.	THE EFFECT OF HABITAT COMPACTION, PATCH SIZE, AND COMMUNICATION DISTANCE	
	ON KILL RATE IN THE BASE CASE.	24
11.	THE EFFECT OF HABITAT COMPACTION AND COMMUNICATION DISTANCE ON THE	
	NUMBER OF OUTBREAKS IN THE BASE CASE	25
12.	THE EFFECT OF HABITAT COMPACTION, PATCH SIZE, COMMUNICATION DISTANCE	
	AND FREE FLIGHT ON KILL RATE IN EXPERIMENT II.	26
13.	THE EFFECT OF HABITAT COMPACTION, PATCH SIZE AND FREE	
	FLIGHT ON THE NUMBER OF OUTBREAKS IN EXPERIMENT II	26
14.	THE EFFECT OF WINTER REPRODUCTIVE RATE ON RESULTS FROM	
	Experiment II.	27
<b>15.</b>	THE EFFECT OF WINTER REPRODUCTIVE RATE ON RESULTS FROM	
	Experiment II.	28
<b>A.1.</b>	Effect of the parameter $\alpha_{_{\!M}}$ on turning angles in free flight	44
A.2.	THE SENSITIVITY OF ATTRACTIVENESS TO THE GRADIENT SENSITIVITY PARAMETERS	45
A.3.		
	GRADIENT AND THE GRADIENT SENSITIVITY PARAMETERS	46
A.4.	SENSITIVITY OF THE RELATIONSHIP BETWEEN PER BEETLE LANDING RATE AND	
	THE NUMBER OF NESTING BEETLES TO HOST RESISTANCE	47
A.5.	SENSITIVITY OF LANDING RATE TO SIX MODEL PARAMETERS	48
<b>A.6.</b>	SENSITIVITY OF PER BEETLE AND PER AREA PHEROMONE PRODUCTION	
	RATES TO HOST RESISTANCE	49
<b>A.7.</b>	SENSITIVITY OF PHEROMONE PRODUCTION TO THREE MODEL PARAMETERS	50
<b>A.8.</b>	SENSITIVITY OF REPRODUCTIVE RATE AND KILL RATE TO HOST RESISTANCE	51
	SENSITIVITY OF REPRODUCTIVE RATE TO THREE MODEL PARAMETERS	
A.10	SENSITIVITY OF REPRODUCTIVE RATE TO THREE MODEL PARAMETERS	53

## **Abstract**

Despite ecological and management importance, little is known about the effect of forest landscape structure on the spread of mountain pine beetles (*Dendroctonus ponderosae* Hopkins). The general prediction from published literature is that forest fragmentation at some scale might slow the spread of infestations. However, mountain pine beetle dispersal ecology is complicated by requirement for attack *en masse* and a pheromone-based communication system that facilitates this aggregation process. One interesting possibility is that infestations might spread more slowly over habitat gaps across which beetles cannot communicate. To investigate this possibility, we develop an individual-based model of mountain pine beetle dispersal, aggregation and attack, and perform simulation experiments to explore the effects of habitat patch size, habitat compaction, communication distance, and flight behaviour on the spread rate and final extent of infestations. The model is based on a spatially explicit mountain pine beetle model developed by Powell *et al.* (1996, *Mathematical elements of attack risk analysis for mountain pine beetles. Journal of Theoretical Biology* 204(4): 601-620), but differs from that model on several points.

Increasing the distance between patches does slow infestation spread, though not as much as expected. When we remove the tendency of beetles to fly for some period before becoming receptive to pheromones (free flight), patch size becomes important and spread rate only varies with habitat compaction when patch size is small. At face value, the prediction is that beetles will be somewhat sensitive to forest fragmentation at small scales, but insensitive to variation in patch size because of their free flight behaviour. This result is subject to uncertainty about model form and parameterization. More robust aspects of this work include a review of the potential effects of relatively small-scale spatial structure on aggregating organisms, a comparison of diffusion and individual-based modeling approaches, and consideration of some assumptions underlying this and preceding mountain pine beetle models.

## Résumé

Malgré son importance sur les plans de l'écologie et de la gestion, on connaît très peu l'effet de la structure du paysage forestier sur la dispersion du dendroctone du pin ponderosa (*Dendroctonus ponderosae* Hopkins). La prévision générale se dégageant de différentes publications est que la fragmentation de la forêt à une certaine échelle pourrait ralentir l'infestation. Cependant, l'écologie de la dispersion du dendroctone du pin ponderosa est compliquée par l'exigence d'une attaque massive et d'un système de communication fondé sur les phéromones qui facilite le processus d'agrégation. Une possibilité intéressante est que les infestations pourraient se répandre plus lentement en cas d'écarts entre les habitats qui empêchent les scolytes de communiquer. Afin d'étudier cette possibilité, nous mettons au point un modèle individualisé de dispersion, d'agrégation et d'attaque du dendroctone du pin ponderosa, et nous effectuons des expériences de simulation dans le but d'explorer les effets de la taille de la parcelle d'habitat, de la compacité de l'habitat, de la distance de communication et du comportement de vol sur la rapidité de propagation et l'étendue finale des infestations. Ce modèle repose sur un modèle spatialement explicite du dendroctone du pin ponderosa mis au point par Powell et al. (1996, *Mathematical elements of attack risk analysis for mountain pine beetles. Journal of Theoretical Biology* 204(4): 601-620), mais il en diffère à plusieurs égards.

L'augmentation de la distance entre les parcelles ralentit la vitesse d'infestation, mais pas autant que l'on pensait. Lorsque nous éliminons la tendance des dendroctones à voler pendant un certain temps avant d'être réceptifs aux phéromones (vol libre), la taille de la parcelle devient un facteur important et la vitesse de propagation varie uniquement en fonction de la compacité de l'habitat si la parcelle est de petite

taille. À l'origine, la prévision est que les insectes seront sensibles en quelle que sorte à la fragmentation de la forêt à petite échelle, mais qu'ils seront insensibles à la variation de la taille de la parcelle en raison de leur comportement en vol libre. Ce résultat dépend de l'incertitude liée à la forme et au paramétrage du modèle. Les aspects plus solides de ces travaux incluent une révision des effets possibles d'une structure spatiale relativement petite sur les organismes qui se regroupent, une comparaison des approches de modélisation individualisée et de la diffusion, ainsi qu'une étude de certaines hypothèses sur lesquelles reposent ce modèle et les modèles antérieurs d'étude du dendroctone du pin ponderosa.

## 1.0 Introduction

#### 1.1 Management context and motivation

Mountain pine beetles (*Dendroctonus ponderosae* Hopkins) are a major agent of lodgepole pine (*Pinus contorta* Dougl. ex Loud. Var. latifolia Engelm.) mortality, and have therefore been a subject of intense research and management effort for over 100 years (Amman and Logan 1998). Population models (Berryman 1978; Clark et al. 1979; Thompson et al. 1981; Raffa and Berryman 1986; Mawby et al. 1989; Safranyik et al. 1999) and more general understanding of system dynamics (Ludwig et al. 1997) predict that outbreaks should be easier to prevent than to control. A long history of failed or dubiously successful control efforts (Wood et al. 1985; Amman and Logan 1998) is consistent with the theory, prompting calls for more proactive management (Amman and Safranyik 1985; Amman and Logan 1998; Samman and Logan 2000).

Lodgepole pine trees become susceptible at 60 to 80 years of age. Thus, one way to prevent mountain pine beetle outbreaks is to remove older pine from the forest (Amman et al. 1984; Amman and Safranyik 1985; Maclauchlan and Brooks 1994). Older forests tend to have high biodiversity, wildlife, recreation and other non-timber values, so a strategy to reduce the susceptibility of a landscape to beetle outbreaks while retaining old forest may best meet societal goals.

Some promise for slowing beetles while retaining pine lies in the consideration of landscape pattern and spatial heterogeneity. For instance, increasing the heterogeneity of landscapes may decrease the rate of spread and extent of disease outbreaks and disturbances (Turner et al. 1989; Rodriguez and Torres-Sorando 2001), decrease the survival and reproductive rates of some organisms (Simberloff 1988; Cantrell and Cosner 1991; Saunders et al. 1991; Andren 1994; Bender et al. 1998; Heibeler 2000), and globally stabilize locally unstable population dynamics (Hastings 1977; May 1978; Reeve 1988; Taylor 1990). In forests, the juxtaposition of stands of different age classes can reduce overall landscape flammability (Franklin and Forman 1987; Turner and Romme 1994; Turner et al. 1999), and isolated forest patches suffer less damage from the eastern spruce budworm (*Choristoneura fumiferana*) (Cappuccino et al. 1998). The effect of landscape pattern depends on complex interactions between a landscape and the habitat preferences and dispersal ecology of individual species (Saunders et al. 1991; Andren 1994; Coulson et al. 1999). For specific predictions it is necessary to consider some details of mountain pine beetle ecology.

## 1.2 Terminology

Throughout this paper we use concepts and terms from the field of landscape ecology (e.g. landscape, patch, spatial heterogeneity), and from population ecology or entomology (e.g. infestation, outbreak, aggregation). Terms common in these fields may not be familiar to a wider audience, and even within a field some confusion may arise. We also have coined a few terms specifically for use in this paper (e.g. free flight, communication distance, habitat compaction). Concepts and terms specific to this paper are explained as they arise throughout this text, and are also included in Table 1b for reference. Concepts and terms common to either landscape ecology or population ecology may not be defined in the text, but are explained with references in Table 1a,b for the convenience of readers.

## 1.3 Expectations

#### 1.3.1 Effect of pattern in general

First, we are concerned with a narrow range of effects arising from variation in the spatial arrangement of habitat. Changing forest pattern can significantly affect radiation fluxes, wind patterns, water fluxes,

predator abundance, and other aspects of the ecosystem (Saunders 1991) that might in turn affect mountain pine beetles. We do not consider such indirect effects.

Models of simple epidemics with removal (SIR – Susceptible/Infective/Removed) are standard in epidemiology (see Capasso 1993 for overview). In these models, disease is transmitted from infective to susceptible individuals, which then become infective. After some period, infective individuals either die, or recover and become resistant to the disease. Since mountain pine beetles kill their hosts as they spread, results from SIR models may be relevant.

Bolker (1999) examined the effect of host spatial heterogeneity, or patchiness, on the spread of an SIR type disease through plant populations. He used analytic and stochastic simulation models to demonstrate that clustering hosts in space can increase both the spread rate and the final size of epidemics, while even spacing of host plants may decrease rate of disease spread (Figure 1a). This result is consistent with previous theoretical work in epidemiology, and the effect has also been observed in several experimental studies (Bolker 1999).

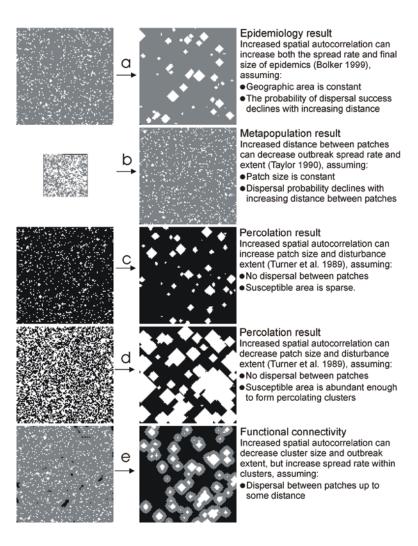


Figure 1: The potential effects of changing habitat configuration depend on organism dispersal ability, habitat abundance, and whether total geographic extent is allowed to vary. Suitable habitat (or susceptible area, for consideration of non-biological disturbances) is white, unsuitable habitat within the dispersal range of an organism is grey, and unsuitable non-traversable area is black.

A central result in metapopulation theory is that increasing the distance between sub-populations can allow regional stability of locally unstable predator-prey interactions by limiting the extent of local population oscillations or outbreaks (Taylor 1990) (Figure 1b). The different predictions from metapopulation and epidemiological work are partly due to different reference conditions. The epidemiologists assume that average initial transmission rate, or the initial ratio of hosts to non-hosts, is constant, so their finding is that epidemics spread faster when hosts are aggregated than when hosts are dispersed randomly over the same geographical area (Figure 1a). In contrast, the metapopulation studies are concerned with increasing the isolation of (or distance between) host patches, so they hold patch size constant and increase the distance between hosts or, effectively, the geographic area across which hosts are dispersed. Increasing the distance between patches will tend to slow the spread of organisms, given constant patch size (Figure 1b). If geographic area is constant, the average distance between the edges of round patches can only be increased by making patches larger, which increases spread rate.

The epidemiological and metapopulation models discussed so far assume that dispersal may occur between patches. In the other extreme, classic percolation models assume that agents cannot spread across gaps between habitat patches, so the total extent of an outbreak or disturbance is absolutely constrained by the size of habitat patches (Turner et al. 1989). In these models, we observe a threshold effect; at intermediate habitat abundance, small increases in habitat can dramatically increase disturbance extent by linking large patches. Percolation models also emphasize that the effect of increased patchiness should depend on the relative abundance of habitat - if habitat is abundant, increased clumping may decrease average patch size and outbreak extent isolating patches from one another (Figure 1d). In contrast, if habitat is sparse, increased clumping may increase patch size and outbreak extent (Figure 1c). Empirical studies of birds and mammals support the general conclusion that pattern effects depend on the relative abundance of habitat (Andren 1994).

In most situations, the dispersal reality is probably somewhere between two extremes. No organisms can traverse infinite distances, but most can travel some distance through inhospitable area. Thus, the "functional connectivity" of habitat is less than infinite, and more than the physical connectivity of habitat patches (Tischendorf and Fahrig 2000). Keitt et al. (1997) have highlighted that clusters of habitat patches separated by less than dispersal distance may be linked into percolation clusters by dispersal. Decreasing average cluster size should decrease the average size of outbreaks (Figure 1e). Within clusters, increasing patch size might increase spread rate as predicted by epidemiological models.

#### 1.3.2 Consequences of pheromone mediated aggregation

Pine resists beetle attack by secreting resin that physically impedes progress, interferes with pheromone signals, seals the living cells from infection by fungus, and drowns eggs and larvae. [For overviews of mountain pine beetle biology see Amman (1978), Amman et al. (1984), Safranyik (1989), and Samman and Logan (2000)]. To overcome host defenses, beetles must attack in large groups. To facilitate mass attack, beetles disperse synchronously over a period of two to three weeks, and use a system of at least five different pheromones to attract others to areas where attack is insufficient, and later repel from areas that are full (see Borden et al. 1987 for review). In entomological literature the process is termed pheromone-mediated aggregation (and attractive pheromones are aggregative). In a more general context, Turchin (1998) distinguishes between aggregative behaviour, where organisms orient toward one another. We note that from this perspective, mountain pine beetle dispersal is functionally more like congregation than aggregation. Throughout this document we use the term aggregation to remain consistent with entomological literature.

Discussions of functional connectivity often focus on dispersal ability as the limiting factor (Keitt et al. 1997; Tischendorf and Fahrig 2000). However, if aggregation is required for success then landscapes across which organisms cannot effectively aggregate will be fragmented, whether or not individuals can move from patch to patch. For mountain pine beetles, functional connectivity may depend on the scale

of pheromone communication, and fragmentation at the scale of pheromone communication may alter the rate and extent of infestation spread. Others have also suggested that attraction between conspecifics might affect population dynamics (Smith and Peacock 1990; Turchin 1989).

#### 1.3.3 Consequences of free flight

The spatial aspect of mountain pine beetle population ecology has intrigued other authors, and other spatially explicit beetle models have been built. Most have taken a continuum reaction-diffusion approach, where movement is approximated by a simple diffusion or advection-diffusion process, written as a partial differential equation (Polymenopoulos and Long 1990; Turchin and Thoeny 1993; Logan et al. 1998). The model most applicable here is by Powell, Logan, Bentz and others (referred to as the MPBpde) (Powell et al. 1996; White and Powell 1997; Logan et al. 1998; Powell et al. 1998; White and Powell 1998; Powell et al. 1999; Biesinger et al. 2000; Powell et al. 2000 – for simplicity, this group of citations will be referred to collectively as Powell et al. through the remainder of this document). The MPBpde is the starting point for our model.

The diffusion approach has several advantages, including generality, flexibility, and analytical tractability (Okubo 1980; Turchin 1991, 1998). The mathematics are well understood, and behavioural assumptions are clear and explicit. However, not all movement behaviours can be handled with a diffusion approach. For example, many bark beetle species require some period of exercise, or "free flight", before they become responsive to pheromones (Borden et al. 1986). Field studies of *Ips typographus* suggest that free flight periods may significantly alter dispersal patterns (Helland 1984; Helland 1989), and indirect evidence for the effect on mountain pine beetle dispersal is given by the fact that some beetles fly past nearby attractive sources (Safranyik et al. 1989), and that infestations tend to spread before all the hosts within an infested area are depleted (Mitchell and Preisler 1991; Borden 1993). In diffusion approximations, individuals are considered in aggregate. Unless all beetles emerge at the same time, one cannot model changes in behaviour that depend on time since emergence with a diffusion approach. Since mountain pine beetles emerge throughout the flight period, a diffusion approximation cannot capture the free flight behaviour. We wish to know whether adding free flight fundamentally alters the results of interest.

Continuum reaction-diffusion models are just one among a range of options for modeling movement (Turchin 1998). At the other extreme, in individual-based models organisms move independently of one another. The individual-based approach allows more flexible movement rules, and the implementation of these rules is largely intuitive. However, individual-based models are often difficult to analyze and communicate (Grimm et al. 1999). Greater flexibility is not always a virtue, making models difficult to compare, and less conducive to general insight (Turchin 1998; Grimm 1999; Grimm et al. 1999). Neither approach is perfect, and the best choice depends on the question. Here, we are interested in whether free flight behaviour alters the effect of landscape pattern on mountain pine beetles. Since free flight behaviour cannot be modeled with a diffusion approximation we take an individual-based approach.

## 1.4 Overview and objectives

Our overall objective is to understand the effect of habitat distribution on the dispersal and success of mountain pine beetles. The key hypothesis is that relatively short gaps between habitat patches might disrupt mountain pine beetle and reduce success because pheromone-based communication is important for success, and beetles communicate effectively across shorter distances than they fly. We refer to the maximum distance across which pheromones effectively spread and beetles effectively communicate as "communication distance". To explore this hypothesis, we first develop a simulation model of mountain pine beetle dispersal, aggregation, and success, building on the work of Powell et al. We simulate a set of landscape patterns that vary in the size of habitat patches (w), and the area across which habitat is dis-

persed (habitat compaction, p). Next, we ask how landscape pattern affects beetles in a set of simulation experiments. We look at the effect of pattern on beetles given the best available estimates of all model parameters. We then consider the sensitivity of model behaviour to two key parameters – the tendency of beetles to fly for some distance before becoming receptive to pheromones, and the over-winter reproductive rate of beetles.

## 2.0 Methods

#### 2.1 The model

The simulation model is based on review of knowledge about mountain pine beetle dispersal ecology, and existing beetle models. The synthesis and perspective gained through model development is a key result of this study, so model details are presented in the results section of this paper. However, it is useful to define spatial extent and resolution immediately. These basic decisions about model structure set the context for experimental design and model form decisions to follow.

First, average beetle dispersal distance under a canopy is less than 1 km, and most beetles seem to travel less than 3 km (Safranyik et al. 1989; Turchin and Thoeny 1993). At the other end of the scale, experience with baited pheromone traps suggests that pheromone plumes effectively extend over distances of 50-100 m (Borden, J. Chief Scientific Officer, Pherotech International Inc., 7572 Progress Way, Delta, BC, V4G 1E9). By these indications, our simulated experimental landscapes should be over 3 km in extent (to avoid excessive edge effects in beetle dispersal), with a grid-cell resolution of at least  $50 \times 50$  m (to resolve pheromone dynamics). We chose a spatial extent of  $5 \times 5$  km, or 2500 ha, and grid cell resolution of  $50 \times 50$  m, or 0.25 ha cell<sup>-1</sup>. Running all simulations on a square landscape of  $5 \times 5$  km, resolved at 0.25 ha cell<sup>-1</sup>, gives a total of 10,000 cells per landscape.

## 2.2 Experimental design and analysis

#### 2.2.1 Pattern generation and experimental design

Binary landscapes were created using a spreading algorithm modified from a simple fire model (Fall 1998). Habitat patches are placed at random, subject to the constraint that new patches cannot start within old ones. Patch sizes are chosen from a Weibull distribution. Each patch spreads from a starting cell up to the chosen size, and the complexity of patch shape depends on the number of neighbours to which each cell spreads in each step. New patches do not spread over old ones, but spreading may continue until patches adjoin. Note that because patches may adjoin (fuse) the actual patch size distribution may differ from the target (Weibull) patch size distribution. Patches are initiated and spread sequentially until the proportion of habitat within the landscape reaches a predetermined level. The (target) distribution of patch size is controlled directly by the Weibull scale ( $\theta$ ) and shape ( $\gamma$ ) parameters. The probability of spreading to a given number of neighbours is normally distributed. We further constrained patterns by insisting that the first patch initiate in the centre of each landscape.

Landscape fragmentation, or, for the purposes of this discussion, the average minimum distance between patches, can be affected by changing patch size, changing the amount of habitat within a given area, or changing the area occupied by a given amount of habitat (habitat compaction). We are interested here in the effects of spatial arrangement, rather than quantity, so we kept total habitat supply constant at 5% of the landscape (i.e. 250 ha out of 5000 ha). We used a multi-factorial experimental design, with four levels of average patch size (w) from small (0.39 ha) to large (14.2 ha), four levels of habitat compaction (p) from 10 to 40 % of the landscape at intervals of 10 %, and five replicates of each resulting pattern type (Figure 2). Parameter values were chosen so the range of fragmentation scales (measured as average minimum distance between patches) spans the range of pheromone communication scales (<50 m to >200 m).

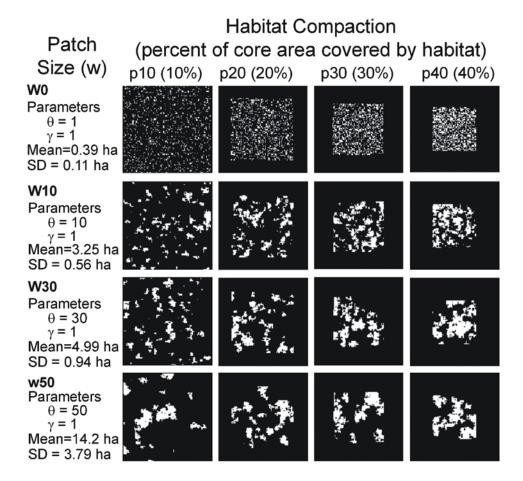


Figure 2: Experimental design scheme showing variation in patch size and patch compaction. Five percent (250 ha) of each landscape is habitat. Weibull scale parameter (θ) is varied and the shape (γ) of the distribution remains constant. The mean and standard deviation (SD) are given for the case where habitat compaction is low (p10). Realized patch size will increase with habitat compaction as close patches merge. There are five replicates of each pattern, one of which is shown here. Communication distance and other experimental factors are repeated within pattern in a split plot experimental design.

#### 2.2.2 Analysis Procedure

The output from each model run is a 25-year time series of area killed per year. Area killed ranges from 0 to 250 ha or 100%. Output curves are characteristically sigmoidal; if conditions allow mountain pine beetles to overcome host defenses then outbreak extent increases exponentially until beetle populations are limited by habitat supply. We fit a three-parameter Weibull-type curve (from Ratkowsky 1990) to each output series (kill vs. time), and then treat the estimated coefficients of each regression as multiple response variables that can be analyzed using standard regression techniques (Cook and Ware 1983). The Weibull-type curve is:

$$y = a_w (1 - e^{-e^{-g_w} X^{d_w}})$$

Increasing either  $g_w$  or  $d_w$  increases the slope of the curve (Figure 3). We analyze the maximum area killed after 25 years (kill<sub>s</sub>) rather than the estimated asymptote  $(a_w)$  because curves that are low but slowly

increasing at the end of the simulation do not provide enough information for a meaningful asymptote estimate. Scale  $(a_w)$  is fixed at 100%. Note that rigorous model selection is not a great concern, as the models are used to describe curves, rather than to predict. If the fit is adequate then comparison of the curves should be consistent regardless of the model used (Potvin et al. 1990).

The parameters  $g_w$  and  $d_w$  are distributed so that a continuous statistical modeling approach is appropriate. The maximum area killed, in contrast, is bimodally distributed. In 15 % of runs, beetles kill less than 20% of habitat over 25 years. Otherwise, beetles kill more than 90% of habitat. To avoid violating regression assumptions we converted this area killed response to a binomial variable by classifying all cases as either "outbreak" (>90% of available habitat killed) or "non-outbreak" (< 90% killed). Because beetles always kill <20% or >90% of available habitat it makes no difference which threshold between 20 and 90% we choose to distinguish outbreak from non-outbreak cases.

We varied communication distance (the maximum distance across which pheromones effectively spread and beetles effectively communicate) and other model parameters within each instance of pattern, so the experiments have a repeated measure or split-block design. The two kill rate responses were analyzed using mixed models to avoid violating the restrictive circularity assumption of randomized block or split-plot ANOVA (von Ende 2001). Patch size and shape are between-subject factors, and all other experimental factors are repeated within pattern. Unstructured covariance matrices were assumed, and all response values were log transformed to equalize variance between groups.

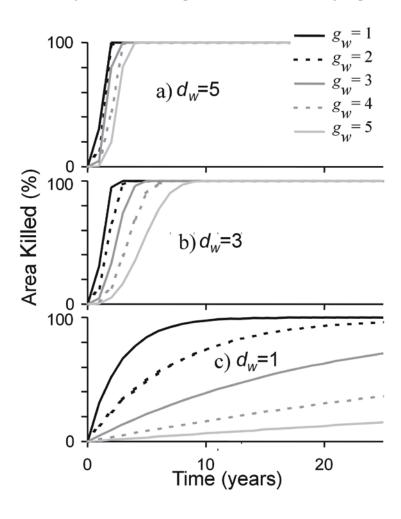


Figure 3: Sensitivity of Weibull-type curve to shape  $(g_w \text{ and } d_w)$  parameters.

Analyzing the binary outbreak data was more difficult. In general, levels of communication distance are not independent; if beetles can outbreak when communication distance is short they are nearly certain to outbreak at larger communication distances. However, lack of variation in response at some levels of communication distance prevents the estimation of more appropriate covariance matrices, leaving us in the ironic position of being unable to analyze the data properly because the experimental effects are too strong. Being unable to fit an appropriate repeated measures model, we pooled the three repeated measures into one response variable by counting the number of outbreaks that occurred on that pattern across all levels of communication distance. We tested for effects of patch size and habitat compaction on this overall chance of an outbreak by fitting a mixed model, assuming the random variable is Poisson distributed, and specifying a log link function (Agresti 1996). Sampling zeros and lack of variation at some levels were corrected by adding a small amount (0.001) to each outbreak count.

Post-hoc comparisons were challenging due to the large number of possible pair-wise comparisons (4560 in the base experiment with four levels of patch size, four levels of habitat compaction, three levels of communication distance, and two response variables). We compared each group mean to the maximum and minimum group means in the experiment to see which kill rates were significantly greater than "low" or significantly less than "high". Although we only examined a subset, all pair-wise comparisons are adjusted for the full set of comparisons using the Tukey-Kramer method.

Data manipulations and non-linear regressions were done with S+ using the NLREGB function (Insightful Corp. 2001, S-Plus V.6.1.Seattle, WA). Continuous mixed models were fit using the MIXED procedure in SAS, and multinomial mixed models were fit using the GENMOD procedure (SAS Institute Inc. 2000, SAS V.8.1, Cary, NC).

## 3.0 Experiments and results

#### 3.1. The model

The annual cycle of the mountain pine beetle can be broadly divided into two parts: the flight period, consisting of several weeks when beetles emerge, disperse, and attack new hosts; and the rest of the year, when beetles feed, mate, and reproduce under the bark of host trees. We describe the model of mountain pine beetle dispersal within a single flight period, then describe how the model is extended over multiple years.

#### 3.1.1. Notation

State variables (e.g., the number of flying beetles, F) in this model vary across space and time, so the value of variable F at location (i,j) and time t is  $F_{i,j,t}$ . For convenience, in most cases we drop the subscript notation on state variables, which are denoted by capital letters. Model parameters are aspatial constants unless otherwise noted. Probabilities are denoted using P().

#### 3.1.2. Conceptual overview of the flight model

Each individual beetle can be in one of five possible modes: *waiting* to emerge, flying and unreceptive to pheromones (*uflying*), flying and receptive to pheromones (*rflying*), *nesting* after landing, or *dead*. Nesting beetles are beetles that have committed to staying at a location; once a beetle is nesting it does not return to flying. Each beetle also has a location, and flying beetles may remain within the current cell or move to one of eight neighbouring locations on the square lattice representing the area under consideration. Nesting beetles produce pheromones, to which receptive flying beetles respond. See Figure 4 for a schematic model overview of the flight model.

Following Powell et al., the change in pheromone concentration (A) over space and time is the sum of production, diffusion, and decay (or loss through the canopy). Like Powell et al., we also consider the resistance of forest R, which varies between 0 and 1, but in our model R is static over the flight period. We have added a second forest variable K to explicitly represent carrying capacity for beetles at each location. By making capacity for beetles explicit, rather than an emergent property of resin depletion, we can ensure areas with little resistance (and few trees) become repulsive to beetles at some point, while areas with high resistance (and many trees) do not become repulsive before resistance is depleted.

#### 3.1.3. Mountain pine beetle mode changes

The state of beetle k at time t is given by its mode  $\Omega_{k,t}$  (waiting, uflying, rflying, nesting, or dead) and location Lock,t (row and column). One mode transition per beetle may occur each time-step. The beetle time-step,  $\Delta t$ , is the temporal resolution of the beetle simulation.  $\Delta t$  must be set small enough that beetles land or die at a rate of less than 100% per  $\Delta t$ , and travel at a rate of less than 1 grid cell per  $\Delta t$ . We consider an appropriate value for  $\Delta t$  in section 3.1.11, once death, landing and movement rate parameters have been defined. The conditional probability of transition to mode  $\Omega_{k,t}$  given that the beetle was in mode  $\Omega_{k,t}$ - $\Delta t$  at time t- $\Delta t$  is written as  $P(\Omega_{k,t} \mid \Omega_{k,t-\Delta t})$ . All but ten mode transition probabilities are zero (Table 2). The probability of a beetle remaining in the same mode is one minus the sum of the probabilities for mode change.

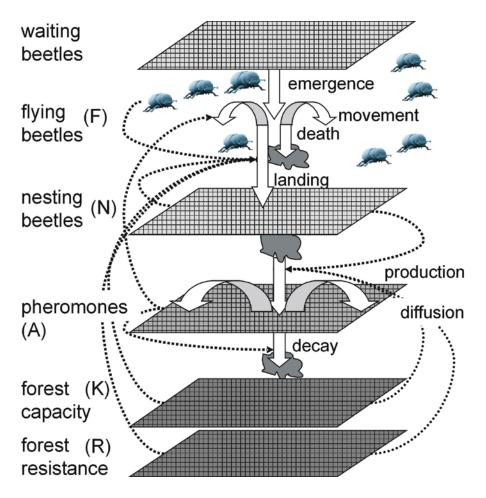


Figure 4: Summary of model within the flight period. Continuous state variables are represented by grids, while individually modeled flying beetles are drawn separately. Descriptions of each state variable are given on the left hand side, with variable names following in brackets. Material flows of pheromones and beetles are indicated by white block arrows. Flows can either be between states (beetles only, vertical arrows), between locations within the same state (curved arrows), or between the modeling system and the external world (vertical arrows beginning or ending in clouds). Flow processes are labeled on the right hand side. Black dotted arrows indicate the effects of state variables on flow rates.

Following Powell et al., we assume that beetles emerge at a uniform rate over the flight period. In truth, emergence patterns depend on temperature patterns over the flight-period and the preceding year, with typically fewer beetles emerging near the beginning of the flight period (Bentz et al. 1991; Safranyik and Linton 1993). The probability of a waiting beetle emerging at a particular time is the inverse of the total flight period:  $P(\Omega_{k,t} = uflying \mid \Omega_{k,t-\Delta t} = waiting) = \Delta t/p_f$ . We assume that waiting beetles represent the number of emergers, so no mortality is applied to this mode. Also following Powell et al., we assume a constant flying death rate:  $P(\Omega_{k,t} = dead \mid \Omega_{k,t-\Delta t} = uflying \text{ or } rflying) = \Delta t\omega_1$ .

To include the free flight mechanism, we follow Helland et al. (1984) and assume that mountain pine beetles are unreceptive to pheromones upon emergence, and unreceptive beetles (*uflying*) become receptive (*rflying*) at a constant rate:  $P(\Omega_{k,t} = rflying \mid \Omega_{k,t-\Delta t} = uflying) = (1 - \Delta t\omega_1)\Delta t r_f$ .

Unreceptive beetles neither land nor respond to chemical cues. Receptive beetles may land with a certain probability, which we discuss in detail in the next section. For now, it suffices that  $P(\Omega_{k,t} = nesting \mid \Omega_{k,t-\Delta t} = rflying) = \Delta t \ landing \ rate_{i,j,t}$ , which depends on the dynamic characteristics of the beetle location (i,j). In reality, beetles land at night when the temperature falls, and return to

flying the next day. We skip over these night-resting periods, and allow beetles to fly continuously through the flight period. Once a beetle has landed it does not return to flying, and nesting death is included in the assessment of over-winter survival and success, so  $P(\Omega_{k,t} = nesting \mid \Omega_{k,t-\Delta t} = nesting) = 1$ . Clearly, dead beetles remain so.

#### 3.1.4. Mountain pine beetle movement

If flying beetles neither die nor land in a time step, they may make one move. The probability of a beetle moving to a location at time t given its state  $\Omega_{k,t}$  (where  $\Omega_{k,t} = uflying$  or rflying) and its location at time t- $\Delta t$  is denoted  $P(Loc_{k,t} = (i+m, j+n) \mid \Omega_{k,t} \& Loc_{k,t-\Delta t} = (i,j))$  where i and j are the row and column location of the beetle at time t- $\Delta t$ , and n and m are integers belonging to the set  $\{-1,0,1\}$ .

Scolytids tend to fly down or across wind until they encounter a pheromone plume, and follow plumes upwind to their source (Choudhury and Kennedy 1980; Byers 1988; Safranyik et al. 1989). Also, in the absence of pheromones beetles preferentially select some hosts based on bole size or host volatiles (Gara et al. 1984). We do not include these complications in the dispersal model. However, Zollner and Lima (1999) have shown that straight or nearly straight search strategies are more efficient than purely random ones. To ensure that search efficiency of beetles is not excessively underestimated, we assume that unreceptive beetles follow a correlated random walk search pattern:

$$P\{Loc_{k,t} = (i+m, j+n) \mid \Omega_{k,t} = uflying \& Loc_{k,t-\Delta t} = (i, j)\} = \frac{e^{\frac{-(1-\cos(\alpha_{k,i+m,j+n,t}))}{(1-\cos(\alpha_m))}}}{\sum_{g=-1}^{1} \sum_{h=-1}^{1} e^{\frac{-(1-\cos(\alpha_{k,i+m,j+n,t}))}{(1-\cos(\alpha_m))}}}$$
(1)

 $\alpha_{k,i+m,j+n,t}$  is the angle between the last move and the potential move (proposed turning angle). The quantity  $(I-cos(\alpha_{k,i+m,j+n,t}))$  is exponentially distributed with the parameter  $\alpha_m$ , which is approximately equal to the average turning angle when  $\alpha_m$  is small (~<60°). Average turning angle does not continue to increase indefinitely with  $\alpha_m$  because the finite range of possible turning angles truncates the exponential distribution. See Figure A.1 for the form of Equation 1.

If there is no difference in pheromone concentration between the current location of a beetle and any of the eight neighbouring cells ( $\Delta A = 0$ ), receptive beetles move like unreceptive beetles (Equation 1). Otherwise, receptive beetles respond to pheromone gradients, or local differences in pheromone concentrations over space. The probability of movement to each location is the relative attractiveness of that location (Figures A.2 and A.3):

$$P(Loc_{k,t} = (i+m, j+m) \mid \Omega_{k,t} = rflying \& Loc_{k,t-\Delta t} = (i, j) \& \Delta A \neq 0) = \frac{Attractiveness_{i+m, j+n,t}}{\sum\limits_{g=-l, h=-1}^{l} Attractiveness_{i+g, j+h,t}}$$

where 
$$\Delta A = \sum_{g=-1}^{1} \sum_{h=-1}^{1} |A_{i,j,t} - A_{i+g,j+h,t}|$$
 (2)

$$Attractiveness_{i+m,j+n,t} = \left(A_{i+m,j+n,t} + \frac{a_r}{a_s - 1}\right)^{a_t} \tag{3}$$

 $A_{i,j,t}$  is the concentration of pheromones at location (i,j) and time t,  $a_r$  is the minimum concentration of pheromones detectable by beetles, and as determines the relative preference of beetles for low-pheromone over no-pheromone areas. The attractiveness parameter,  $a_p$  controls the sensitivity of mountain pine beetles to pheromone gradients. The attractiveness function is designed to make beetles more sensitive to differences in pheromone concentration when pheromone concentrations are low.

Note that beetles interpret pheromones differently in this model than in the MPBpde (Powell et al.). Here, pheromones are always attractive (Equation 3), while in that other model very high pheromone concentrations are unattractive. See the following section for reasoning behind this change.

#### 3.1.5. Landing and pheromone production

Mountain pine beetles facilitate mass attack using a system of at least five semiochemicals that differ in the messages they carry, and the rates at which they are produced and decay. Modeling the mechanics of this system is not practical, so the goal here is an abstraction of the system that adequately reproduces the large-scale behaviour of beetle populations. Before developing the abstraction, a brief review of mountain pine beetle pheromone ecology is in order. Early in attack, females produce the attractive pheromone trans-verbenol (Borden et al. 1987). Myrcene and other volatile monoterpenes are released from resin ducts severed by the attacking beetles. Males attracted by this initial combination of volatiles release multi-functional pheromones (exo-brevicomin and frontalin) that are attractive at low concentrations and repulsive at high concentrations. This multi-functionality may help beetles to avoid overcrowding while still promoting enough aggregation to overcome host resistance.

Later in the attack, females stop producing trans-verbenol while males stop producing exo-brevicomin but continue releasing frontalin. Verbenone, an antiaggregant, is produced by autoxidation of
trans-verbenol and by microorganisms in association with female beetles. In this last phase, high levels
of verbenone and frontalin deter beetles from approaching or landing at the attacked site. The clumped
attack patterns and switching behaviour characteristic of mountain pine beetle could be explained by the
differential decay rates of verbenone and frontalin. Since verbenone photoisomerizes rapidly on exposure
to sunlight (Kostyk et al. 1994), frontalin is likely to diffuse further from the source tree. Beetles would
continue to be attracted to those adjacent areas where verbenone was absent and frontalin was present in
low (attractive) concentrations.

The net effect is that areas become attractive early in attack, and unattractive later. Powell et al. have chosen to model a single pheromone that is attractive at low concentrations and repulsive at high concentrations. In that model the transition from attraction to repulsion is independent of the resistance and the capacity for beetles at a location. Thus, areas with low capacity (or few trees) may never accumulate enough beetle attacks to become repulsive, while areas with high capacity may become repulsive before enough beetles aggregate to overcome host defenses.

Another consequence of Powell et al.'s formulation is that unattractive areas can create barriers to beetle spread, trapping beetles in poor areas. Real beetles are apparently not so constrained; beetles attack trees baited with verbenone (a repellent) and exo-brevicomin (at attractive concentrations) significantly less than trees baited with only exo-brevicomin, but do not attack verbenone-baited trees less than unbaited trees (Shore et al. 1992). Thus, verbenone effectively masks or neutralizes attractive signals, but does not actively deter beetles.

To avoid erecting pheromone barriers and deterring flying beetles from areas that are not yet full, we assume the antiaggregative effects of verbenone and the multi-functional pheromones remain local within each cell. Thus, areas may cease to be attractive, but never become actively unattractive to dispersal. Once a beetle arrives, antiaggregants will deter it from landing if beetle capacity is full.

When population density is low, beetles preferentially attack weakened or diseased trees (Gara et al. 1984; Powell et al. 1996). In the absence of other nesting beetles, we assume flying beetles prefer to land in areas with low resistance to attack, so base landing rate is:

$$r_{base} = r_b (1 - R)^{r_2} (4)$$

R is a value between 0 and 1 that determines the relative resistance of trees to attack,  $r_b$  is the maximum possible landing rate of pioneer beetles, and  $r_2$  controls the selectivity or preference for low resistance over high resistance areas. Landing rate increases as the number of nesting beetles (N)

increases, peaks at some intermediate number of beetles, and declines to zero as the number of nesting beetles reaches capacity (K) (Figures A.4 and A.5):

landing 
$$rate_{i,j,t} = [1 - (\frac{N}{K})^{r_5}][r_{base} + (r_m - r_{base})(1 - e^{-r_1 \frac{N}{K}(1 - R)^{r_3}})]$$
 (5)

 $r_m$  is the maximum possible landing rate,  $r_1$  determines how fast landing rate increases with the number of nesting beetles,  $r_3$  determines how the relative preference of flying beetles for low resistance areas persists as the number of nesting beetles increases, and  $r_5$  determines the sensitivity of flying beetles to crowding.

The long-range attractive effects of trans-verbenol and the two multi-functional pheromones are modeled with a single attractive pheromone, A. Production rate (per beetle) is maximum  $(a_m)$  when nesting beetles are sparse, and decreases to zero as host capacity is reached. Total pheromone production rate (per time) at location i,j is (Figures A.6 and A.7):

pheromone production 
$$rate_{i,j} = Na_m (1 - \frac{1 + e^{a_1 R^{a_6} (R^{a_3} - 1)}}{1 + e^{a_1 R^{a_6} (R^{a_3} - N/K)}})$$
 (6)

The beetle density at which pheromone production reaches one half of maximum increases with resistance, R, and the rate of increase is determined by  $a_3$ . The parameter  $a_1$  determines the steepness of production decline, and  $a_6$  ensures that initial per beetle pheromone production is the maximum unless host resistance is very nearly zero.

We have assumed that flying beetles keep aggregating and landing until areas are full, while nesting beetles stop actively producing aggregating pheromones as soon as host resistance is overcome. While these suppositions remain speculative, it is interesting to note that a conflict between the interests of flying and nesting beetles could account for the multi-functionality of frontalin and exo-brevicomin. In the interest of reducing competition, it makes sense that nesting beetles should put some resources toward producing an "antiaggregation" signal, rather than simply falling silent. However, flying beetles should interpret this signal as an indicator of a secure resource, and aggregate towards it until some threshold concentration is reached. If beetle behaviour is optimal, the concentration at which attractive pheromones become unattractive should correspond to the point at which the cost of competition equals the advantage of security. Finally, the rapid decay rate of verbenone (Kostyk et al. 1994) could explain why this pheromone functions only as an antiaggregant.

#### 3.1.6. Pheromone diffusion and decay

Following Powell et al. (1996) we represent pheromone dynamics using a simple linear diffusion model with a constant decay rate,  $\delta_a$ :

$$\frac{\partial A}{\partial t} = b_a \nabla^2 A + pheromone \ production \ rate_{i,j} - \delta_a A \tag{7}$$

The first term  $(b_a \nabla^2 A)$ , known as the heat equation, is a standard model for approximating the aggregate behaviour of many particles that each move stochastically. The equation is used to describe the spread of heat or the diffusion of dissolved substances through a homogeneous medium. It is also used to model the stochastic (random walk) component of organism movement (Okubo 1980; Okubo and Levin 2001; Turchin 1998), and is the basis for the continuous approach to beetle dispersal modeling taken by Powell et al. in the MPBpde.  $b_a$  is known as the diffusion coefficient, and determines the rate of spread. In a random walk,  $b_a = \lambda 2/2T$ , where  $\lambda$  is step length and T is the period of time between two consecutive

moves.  $\nabla^2$  is the Laplacian operator. In discrete two-dimensional space,  $\nabla^2 A$  (at grid cell x, y) is equal to the sum of differences in pheromone concentration between adjacent locations:

$$\nabla^2 A_{x,y} = (A_{x-1,y} - A_{x,y}) + (A_{x+1,y} - A_{x,y}) + (A_{x,y-1} - A_{x,y}) + (A_{x,y+1} - A_{x,y})$$

#### 3.1.7. Complete model of dispersal, aggregation and attack

In sum, at any point in time, *t*, throughout the flight period, a mountain pine beetle *k* may be in one of five modes. Waiting beetles can emerge, flying beetles can become receptive to pheromones, receptive flying beetles can land (Equations 4 and 5), and all flying beetles can die (Table 2). The population of flying and nesting beetles at each location are F and N, respectively.

$$F_{i,j,t} = \sum_{k=1}^{\text{num beetles at i, j}} (\Omega_{k,t} = \textit{uflying}) + \sum_{k=1}^{\text{num beetles at i, j}} (\Omega_{k,t} = \textit{rflying})$$
(8)

$$N_{i,j,t} = \sum_{k=1}^{\text{num beetles at i, j}} (\Omega_{k,t} = nesting)$$
(9)

At each time-step flying beetles can also change locations. The probability of an unreceptive beetle (*uflying*) moving to any adjacent location depends on the direction of that beetle's flight in the previous time-step (Equation 1). A receptive beetle (*rflying*) will follow a pheromone gradient ( $\nabla A$ ) where such a gradient is present (Equations 2 and 3). Otherwise, it follows the same correlated random walk as unreceptive beetles (Equation 1). Finally, pheromones arise, spread, and decay according to Equations 6 and 7.

The model follows the MPBpde (Powell et al.) in overall structure; nesting beetles produce pheromones to which flying beetles respond, and landing rate, nesting success, and pheromone production depend on forest attributes. We also follow on a number of particulars, including: uniform emergence of beetles over the flight period; constant death rate of flying beetles; and pheromone dynamics modeled by simple diffusion with a constant decay rate.

This model differs from Powell et al. in several ways. First, flying beetles are modeled individually. The correlated random walk followed by unreceptive beetles or receptive beetles in the absence of pheromones (Equation 1) differs somewhat from the simple diffusion of flying beetles in the absence of pheromones in the MPBpde, but that difference is not particularly important. More crucial differences are in how forest is represented (R is static and K is new), how landing rate and pheromone production vary with forest attributes and crowding (Equations 4, 5 and 6), and how flying beetles interpret pheromones (Equations 2 and 3). See the preceding sections for reasoning behind the differences.

#### 3.1.8 Extending the model over multiple years

To extend the model, we consider the attack efficiency of beetles. Provided trees are at least somewhat resistant, beetle success, measured as per capita reproductive rate, should be low when the density of nesting beetles is low, and increase as the number of nesting beetles increases enough to overcome host defenses. As the number of nesting beetles increases even further towards carrying capacity, per capita reproductive rate will decrease due to increasing competition (Safranyik 1999). Furthermore, the point at which host defenses are overcome and survival reaches the maximum should be the point at which pheromone production by nesting beetles declines (Figures A.8 to A.10):

reproduction rate<sub>i,j</sub> = 
$$r_w (1 - \frac{(N/K)^{s_5}}{s_4}) (\frac{1 + e^{s_1 R^{s_6} (R^{s_3} - 1)}}{1 + e^{s_1 R^{s_6} (R^{s_3} - N/K)}})$$
 (10)

The parameters  $s_1$ ,  $s_3$ , and  $s_6$ , control the relationship between reproductive rate and resistance the same way that  $a_1$ ,  $a_3$ , and  $a_6$  control the relationship between pheromone production and resistance.  $r_w$  determines the maximum per capita reproductive rate in the absence of resistance or competition, which we refer to as the winter production rate.  $s_4$  determines the minimum reproductive rate that flying beetles will tolerate (or the reproductive rate at carrying capacity); once an area has reached carrying capacity, flying beetles will always choose to keep searching for less crowded habitat (landing rate goes to zero).  $s_5$  determines the rate at which reproductive success decreases due to crowding.

The number of emerging beetles year n+1 is then (Figures A.8 to A.10):

$$\gamma_{i,j,n+1} = N_n(reproduction \ rate_{i,j,n}) \tag{11}$$

If trees have no resistance, the proportion of trees killed at a location is equal to the proportion of trees attacked. Given the simplifying assumption that beetles fill trees to capacity before moving on, the kill rate is given by N/K. If trees have resistance, then the kill rate will be less than N/K. We assume the decrease is equal to the ratio of realized reproduction rate to maximum reproduction rate. Realized reproduction rate is given by Equation 10, and maximum reproduction rate is the realized rate when R = 0. Dividing through, we get (Figures A.8):

$$kill\ rate_{i,j} = \frac{N}{K} \frac{(1 + e^{s_1 R^{s_6} (R^{s_3} - 1)})}{(1 + e^{s_1 R^{s_6} (R^{a_3} - N/K)})}$$
(12)

As trees are killed, capacity declines by the kill rate:

$$K_{n+1} = K_n(1 - kill \ rate_n) \tag{13}$$

However, assuming that surviving trees fully recover before the next flight period, and all trees within a stand have the same relative resistance, resistance within the stand R remains constant over the course of the outbreak. In truth, trees within a stand may vary widely in their resistance to mountain pine beetles, and resistance also varies over time as trees age (Shrimpton and Thompson 1983; Thomson 1987), weather conditions change (Thomson and Shrimpton 1984; Thomson et al. 1984), or mortality changes stand density (Amman et al. 1988; Amman and Logan 1998).

#### 3.1.9. Base model parameterization

In a field study designed to parameterize the MPBpde, Biesinger et al. (2000) found that the average number of nesting beetles in colonized trees ranged from 567 to 1496 MPB tree<sup>-1</sup>. Stands susceptible to mountain pine beetle attack typically range from between 750 to 1500 stems ha<sup>-1</sup> (Shore and Safranyik 1992; Whitehead et al. 2001). An average stand density of 1000 susceptible stems ha<sup>-1</sup> suggests an average beetle capacity on the order of 1,000,000 MPB ha<sup>-1</sup>. For computational efficiency, we model "individual" groups of 1000 beetles that emerge and move together. This may increase the efficiency of spread at the margins of the infestation, but as long as aggregation of several groups of beetles is required for success the error should be acceptable. To avoid confusion we present all results and parameter values in thousands of MPB, or kMPB. Thus, the average carrying capacity, K<sub>0</sub>, is 1000 kMPB ha<sup>-1</sup>.

Following Powell et al., we assume some variability in forest composition over space so that weak, low capacity areas can provide foci for attack. Beetle capacity is uniformly distributed with a mean of 1000 kMPB ha<sup>-1</sup>, a minimum of 500 kMPB ha<sup>-1</sup>, and a maximum of 1500 kMPB ha<sup>-1</sup>. We also assume that resistance varies normally across the landscape, with a mean of 0.5, and a standard deviation of 0.3. Again, actual resistance among stands is a complex phenomenon dependent on age and composition

of stands, and weather (Amman et al. 1988; Amman and Logan 1998; Shrimpton and Thompson 1983; Thomson 1987; Thomson and Shrimpton 1984; Thomson et al. 1984).

Before adding the complications of chemotaxis, it is worthwhile to consider how beetles move in the absence of guidance. Together, the parameters  $K_o$ ,  $R_o$ ,  $\omega_I$ ,  $r_f$ ,  $\alpha_m$ ,  $r_b$ ,  $r_2$ , and the beetle step rate  $\Delta t$  determine the distance traveled by beetles across a contiguously forested landscape in the absence of pheromones. Following Biesinger et al. (2000), we use a constant death rate ( $\omega_I$ ) of 0.01 fh<sup>-1</sup>. (The base unit of time in this analysis is the flight-hour, fh.) There are approximately 5 fh per day, because beetles only fly during the heat of the day. We do not model the resting period between each flight day; in this model beetles fly continuously through the flight period. The free flight duration of mountain pine beetles is not known, but related bark beetle species require an average of between 30 and 90 minutes flight exercise before becoming receptive to olfactory stimuli (Borden et al. 1986). Lacking better estimates, we assume that beetles become receptive to pheromones at a rate (rf) of 0.65 fh<sup>-1</sup>. Finally, beetles turn an average of angle,  $\alpha_m$ , of 35° at each step. See Figure A.1 for the effect of  $\alpha_m$  on the distribution of turning angles.

In the study from which Biesinger et al. (2000) derived the MPBpde movement parameter estimates, Turchin and Thoeny (1993) found that 50 % of southern pine beetles disperse less than 0.69 km, and 99% disperse less than 3.29 km. A step rate ( $\Delta t$ ) of 0.25 fh cell-1 (where each raster cell is 50 × 50 m or 0.25 ha) and a maximum base landing rate ( $r_b$ ) of 0.2 fh-1 gives an average travel distance of 0.64 km on contiguous habitat, and 1.3 km on a landscape without habitat (but note that the latter may be an underestimate due to edge effects – see Figure 5 for details). 90% of beetles travel less than 1.5 km on contiguous habitat (Figure 5).

17

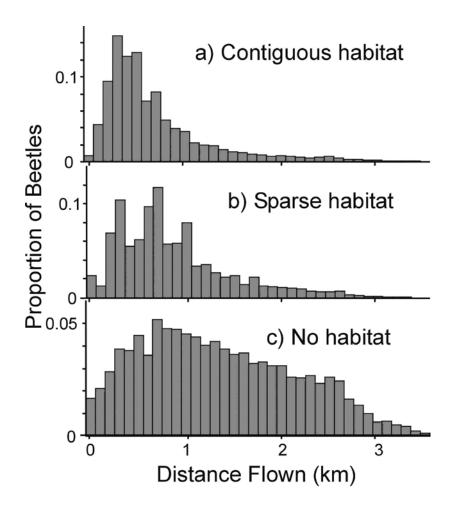


Figure 5: Distances flown by modeled mountain pine beetles in the absence of pheromones on a) contiguous habitat (100% of the landscape suitable for landing) b) sparse habitat (10% of the landscape suitable) and c) no habitat (0% of the landscape suitable). In the first two cases beetles can either die or land, but in the absence of habitat flight distance is limited by death alone. The average flight distance on contiguous habitat is 0.64 km, and the median is 0.48 km. With no habitat the average flight distance is 1.3 km, and the median is 1.2 km. Note that beetles cannot fly more than 3.5 km from their source at the centre pixel because the square landscapes are only 5 km across. The flight distance distributions suggest that when habitat is available edge effects are not severe, but in the absence of habitat they are more so.

Biesinger et al. (2000) estimate that pheromones are produced at a rate of 20  $\mu$ g fh<sup>-1</sup> kMPB<sup>-1</sup>, diffuse at an average of b<sub>a</sub> = 0.648 ha fh<sup>-1</sup>, and decay at an average rate of  $\delta_a$  = 180 fh<sup>-1</sup> in a stand of average openness with a wind speed of 0.6 m s<sup>-1</sup>. We accept these base diffusivity and decay rate estimates, and assume a maximum pheromone production rate, a<sub>m</sub>, of 20  $\mu$ g fh<sup>-1</sup> kMPB<sup>-1</sup>. Following Geiszler et al. (1980), we assume that beetles are sensitive to concentration of pheromone greater than 3×10<sup>-3</sup> ng m<sup>-3</sup>. We further assume chemicals more than 3 m from the ground are lost to the system, so the minimum detectable pheromone concentration is  $a_r$  = 0.1  $\mu$ g ha<sup>-1</sup>. On a grid cell resolution of 0.25 ha, these spread parameters give a radially symmetric pheromone plume shown in Figure 6. The prediction that beetles can only communicate effectively over distances less than 50-100 m is supported by experience with pheromone baited trapping, where baits should be set no more than ~75 m apart to be most effective (Borden, J. personal communication).

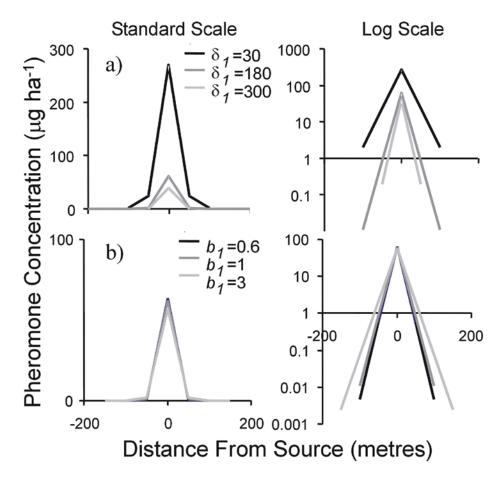


Figure 6: Sensitivity of pheromone plume size and shape to a) pheromone decay rate  $(\delta_1)$  and b) pheromone diffusivity  $(b_1)$ . The number of nesting beetles, N, is 3/5 of carrying capacity (K).

Reproduction, pheromone production, and landing parameters remain to be specified. We first assume that pheromone production declines to zero as reproductive rate increases to maximum (see Figures A.4-A.7 for landing and pheromone production curves). Thus, the parameters that control the reproduction and pheromone production curves are equal  $(a_1 = s_1, a_3 = s_3, \text{ and } a_6 = s_6)$ . Similarly, landing rate declines with reproduction rate, so  $r_5 = s_5$  and per area beetle production rate is maximum when beetles are at capacity (N = K), implying that  $s_4 = s_5 + 1$ . The ratio of emerging to attacking beetles must be at least greater than 1 for the population to increase. We select  $r_m = 3.5$  for the base case.

We chose the pheromone sensitivity parameters  $a_1$  and  $a_s$ , the production parameters  $s_1$ ,  $s_3$ ,  $s_5$ , and  $s_6$ , and the landing parameters  $r_1$  and  $r_3$  so that beetles effectively aggregate, aggregation is important for success, and the beetle success and landing curves look reasonable (Figures A.4-A.10). For a complete description of all parameters and their associated baseline values, see Table 3.

#### 3.1.10 Numerical methods and model implementation

The mixed reaction-diffusion and individual-based modeling approach limits implementation options. Software well equipped to solve partial differential equations is not well designed to represent individuals, and tools helpful for individual-based modeling do not include more advanced mathematical capacities. To resolve this dilemma without excessive programming difficulty, we used an explicit first order forward-Euler method to solve the diffusion equations. This method is less accurate and has more restrictive stability conditions than other methods, but can be implemented without the use of sparse-matrix solvers

or other mathematical tools. As long as the following criterion is met, the method is stable (Sewell 1988):

$$\Delta t_a \le \frac{1/4H}{b_a}$$

 $\Delta t_a$  is the pheromone time-step, H is the grid cell resolution, and  $b_a$  is the diffusion coefficient for pheromones. Accuracy of the solution is not a primary concern because neither our model nor our questions are precise; we care about whether beetles can communicate over 50 or 200 m, not the precise shape of the diffusion curve.

We implemented the model using SELES, a declarative modeling language for spatio-temporal modeling (Fall and Fall 1999; Fall and Fall 2001 — note that the language has been expanded to allow individuals since the original release). Reflective boundary conditions are assumed throughout.

#### 3.1.11 Temporal extent and resolution

Pheromones decay at a rate of 180 fh<sup>-1</sup> and spread at a rate of 0.648 ha fh<sup>-1</sup> (Biesinger et al. 2000), suggesting that pheromone dynamics must be resolved at time step of less than 1/180 fh. In contrast, beetles move at a rate of 1 ha fh<sup>-1</sup>, and land or die at a rate of less than 1 fh<sup>-1</sup>, suggesting that a time-step of 1/4 fh is adequate to resolve beetle dynamics. To speed computation, White and Powell (1998) solve the pheromone equations analytically over the beetle time step. We follow the spirit of this approach, but, lacking the tools required for transformation to Fourier space, we solve the pheromone equation numerically. After each beetle time-step we simulate pheromone dynamics until the chemical landscape is near enough to equilibrium that further simulation causes changes of less than 10-3 µg ha<sup>-1</sup> in chemical concentration, and then proceed to the next beetle time step.

Following Logan et al. (1998) we assume a baseline flight period of 8 days with 5 fh per day, giving a total annual flight period,  $p_f$  of 40 fh in which all beetles disperse. Note that the duration of flight period in any year depends on weather and local conditions, and can vary from 3 days to 3 weeks (Safranyik et al. 1978). We ran each simulation for 25 years, which was long enough for beetles on contiguous habitat to successfully kill all susceptible trees. Beetle dynamics are resolved with a time-step of  $\Delta t = 0.25$  fh. Pheromones dynamics run on a time-step of  $\Delta t_a = 1/500$  fh.

#### 3.1.12 Initial conditions

We begin each run with 1,250,000 successful beetles in the center of each landscape, giving an initial emergence rate of  $1250r_w/p_f = 109.375 \text{ kMPB fh}^{-1}$  from the center pixel.

## 3.2 Clarifying hypotheses

The underlying hypothesis is that beetles are constrained by pheromone connectivity, which is in turn influenced by the pattern of the forest mosaic. That is, infestations should spread efficiently between patches near enough to one another that pheromones can travel from one to the other, and slowly or not at all across larger gaps.

To measure how landscapes are connected by pheromones, we join patches less than communication distance from one another into connected clusters (Keitt et al. 1997). Since beetles spread from the center in all landscapes, we measure only the center cluster (Figure 7). If beetle spread is strictly limited by whether or not beetles can communicate across gaps then the final extent of the infestation should be equal to the center cluster size in each landscape. We analyzed the effect of experimental factors on center cluster size using mixed models (see the method for analyzing  $g_w$  and  $d_w$ ).

In general, the effect of increasing any of habitat compaction (p), patch size (w), and communication distance (or buffer width, d) while holding the others constant is to either increase center cluster size (ccs) or leave it unchanged. Interactions between variables are significant  $(ccs \sim d^*w^*p, n=320, p<0.0001$  – Note: throughout the remainder of this document, the notation "response  $\sim$  factor1\*factor2" followed by a sample size (n) and a p-value indicates that these two factors significantly interact in their effect on the response variable. Sample sizes are for both within and between-subject factors, but all factors except p and w are within-subject so the number of independent samples is less than n. The three possible response variables are abbreviated as: ccs = centre cluster size, kr = kill rate, and op = outbreak probability. The two kill rate parameters  $(g_w$  and  $d_w$ ) are treated as repeated measures of the same response (kill rate -kr). rpt is the factor name associated with the two dependent variables  $(g_w$  and  $d_w$ ). A significant effect of rpt indicates that the two kill rate parameters  $(g_w$  and  $d_w$ ) differ in their response to the experimental factors).

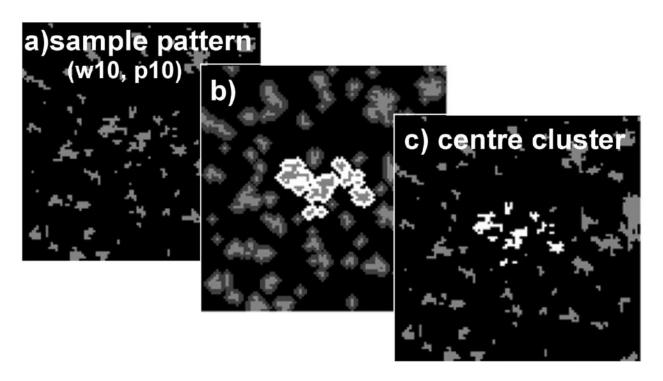


Figure 7: Definition and calculation of center cluster size. Patches nearer together than communication distance (200 m, in this case) are joined to delineate connected "clusters" of habitat (b). The centre cluster consists of the habitat patch at the centre of the landscape, and all habitat patches to which it is connected (c). Center cluster size is the sum of habitat area within this cluster.

When patch size is small (w0) and communication distance is large (d = 200 m), the distance between patches is less than 200 m, so the landscape is connected (center cluster size ~100%) regardless of habitat compaction level (Figure 8). Increasing habitat compaction decreases the distance between patches enough to move the landscape from almost completely unconnected to completely connected when communication distance is intermediate (Figure 8, w0, d = 50 m/100 m). When communication distance is zero, the effect of habitat compaction is insufficient to connect the landscape (Figure 8, w0, d = 0 m).

When patch size is small, the transition from connected to unconnected occurs over a relatively small range of habitat compaction (10-20%), demonstrating the threshold-type behaviour characteristic of percolation networks (Figure 8, w0). As patch size becomes larger (w10 - w50), the distance between patches becomes both larger and more variable. As this happens, differences between communication distance levels tend to diminish (Figure 8). The effect of habitat compaction also becomes more continuous, and the threshold effect disappears.

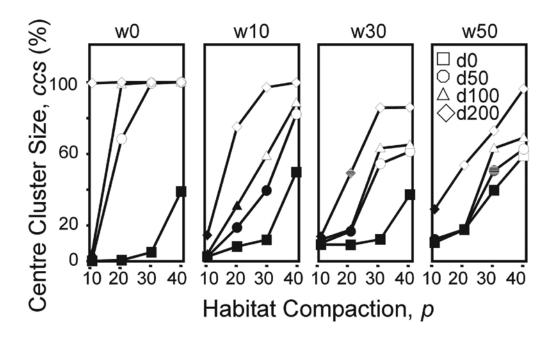


Figure 8: The effect of habitat compaction (area over which habitat is dispersed, *p*), patch size (*w*), and communication distance (buffer width, *d*) on centre cluster size (*ccs*). If infestations can only spread efficiently over habitat gaps across which beetles can communicate, then centre cluster size should predict final infestation extent. These show the experimental effects expected if this hypothesis is correct. Overall, the interaction between all three variables is significant (*p*=0.0001). Open symbols show group means that are not significantly different from the maximum (250 ha), while closed symbols indicate no significant difference from the minimum. Group means marked with stripes are significantly different from both the maximum and the minimum. Each point marks the mean value for five replicate patterns.

#### 3.3 Experiment I - base case

The questions are: What is the effect of patch size and patch compaction on infestation spread rate and extent, and how does pheromone communication distance alter these effects? Is beetle success directly constrained by pheromone connectivity, and can beetles aggregate efficiently only over gaps across which they can communicate?

Pheromone communication distance can be altered by adjusting one of two parameters in the model. All else being equal, decreasing pheromone diffusivity  $(b_a)$  and increasing decay rate  $(\delta_a)$  both decrease the extent of a pheromone plume. Diffusivity has a relatively small effect on plume extent (Figure 9), so we only changed decay rate. To see how the effect of pattern varies with communication distance we vary pheromone decay rate across three levels. d50 is 0-50 m, d100 is 50-100 m, and d200 is 100-200 m (Figure 9). We expect that both outbreak frequency and kill rate  $(g_w, d_w)$  should increase with increasing habitat compaction, communication distance, or patch size. Interactions between variables should be like those in Figure 8.

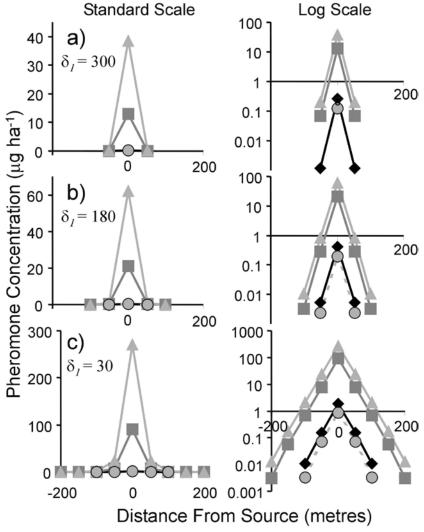


Figure 9: Pheromone plumes used in base experiment. Graphs a) through c) show how changing pheromone decay rate  $(\delta_I)$  alters the relationship between pheromone plume size and shape and the number of nesting beetles, N.  $\blacktriangle$ : N = 0.001\*K,  $\blacksquare$ : N = 0.2\*K,  $\spadesuit$ : N = 0.6\*K,  $\bigcirc$ : N = 0.996\*K. A decay rate of  $\delta_I$  = 300 (a), given a communication distance of 0-50 m (d50),  $\delta_I$  = 180 (b) gives a distance if 50-100 m (d100), and  $\delta_I$  = 30 (c) gives a distance of 100-200 m (d200).

The effect of habitat compaction on kill rate is significant ( $kr \sim p$ , n = 240, p=0.003), and kill rate generally increases with habitat compaction as expected (Figure 10). The interaction between communication distance and the two response variables ( $g_w$ ,  $d_w$ ) is also significant ( $kr \sim d^*rpt$ , n = 240, p<0.0001). Increasing communication distance from 50 to 100 m consistently increases kill rate as expected. However, increasing communication distance from 100 to 200 m only appears to increase kill rate when habitat compaction is low (Figure 10). When habitat compaction is high, kill rate tends to decrease as communication distance increases from 100 to 200 m. This apparent interaction between habitat compaction and communication distance is not significant ( $kr \sim d^*p^*rpt$ , n = 240, p=0.0701), but the trend is consistent over all patch sizes. Kill rate appears to increase slightly with patch size, especially when communication distance is small (d50), but this effect is also not significant ( $kr \sim w$ , n = 240, p=0.077).

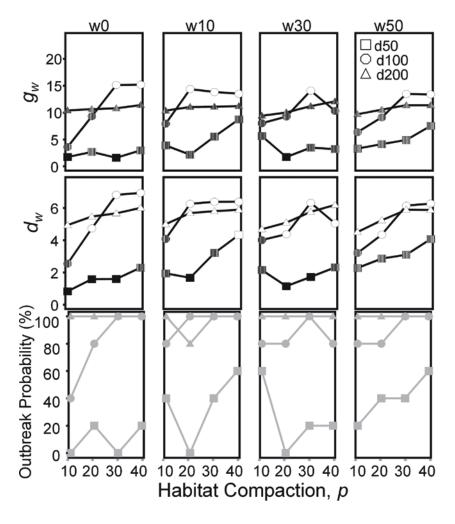


Figure 10: The effect of habitat compaction (p), patch size (w) and communication distance (d) on kill rate  $(g_w, d_w)$  in the base case. Note that the interaction between communication distance and response variable is significant (p<0.0001), as is the effect of habitat compaction (p=0.0003), but the effect of patch size is not. Open symbols show group means that are not significantly different from the maximum, while closed symbols indicate no significant difference from the minimum. Groups marked with stripes are not significantly different from either the maximum or the minimum. Outbreak probabilities are shown for interest, but the data could not be analyzed in this form due to lack of variation when communication is maximum (d200).

In general, trends in outbreak probability mirror trends in kill rate, except that kill rate may vary between cases where outbreak probability is consistently 100% (Figure 10). However, relationships between outbreak frequency and the experimental factors are not significant (Figure 11). Experimental results are summarized in Table 4.

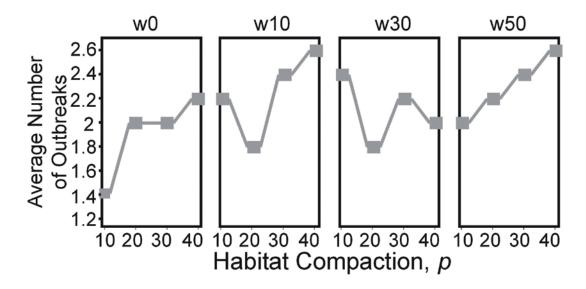


Figure 11: The effect of habitat compaction (*p*) and patch size (*w*) on the average number of outbreaks (*op*) in the base case. Outbreaks are tallied across three levels of communication distance for each pattern, so the maximum number of outbreaks possible is 3, and the minimum is 0. However, no landscape had less than 1 outbreak in this case because outbreaks always occurred when communication distance was large. Patch size (*w*) and habitat compaction (*p*) do not have a significant effect on the average number of outbreaks.

## 3.4 Experiment II - the effect of free flight

The question is: How does the tendency of beetles to fly for some time before becoming responsive to pheromones alter the results from Experiment I? To answer this question, we repeat the Experiment I with no free flight period (r=4) and a reduced pattern set (w0 and w50 only).

We expect that without free flight, center cluster size might be more likely to constrain infestation size because beetles are less likely to fly towards areas without attractants as they do in free flight. Overall, we expect that free flight should be advantageous to beetles.

The effect of free flight on kill rate varies with response variable  $(g_w \text{ or } d_w)$  and communication distance  $(kr \sim r_f^* d^* rpt, \text{ n=240}, \text{ p<0.0001})$ , with response variable  $(g_w \text{ or } d_w)$  and habitat compaction  $(kr \sim r_f^* p^* rpt, \text{ n=240}, \text{ p=0.0486})$ , and with communication distance and patch size  $(kr \sim r_f^* d^*w, \text{ n=240}, \text{ p=0.0053})$  (Figure 12). Removing free flight also increases the frequency of outbreaks overall  $(op \sim r_f^* rpt, \text{ n=20}, \text{ p<0.0001})$  (Figure 13). Essentially, removing free flight eliminates all experimental effects by increasing spread rate and outbreak frequency when communication distance is small, and decreasing spread when communication distance is intermediate and habitat compaction is high (Figure 12).

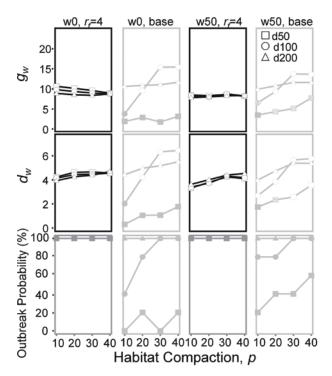


Figure 12: The effect of habitat compaction (p), patch size (w), communication distance (d) and free flight  $(r_f)$  on kill rate  $(g_w, d_w)$  in Experiment II. Note that the interaction between patch size, communication distance, and free flight is significant (p = 0.0053), as is the interaction between free flight, response variable and habitat compaction (p = 0.0486) and the interaction between free flight, response variable and communication distance (p < 0.0001). Grey panels are repeated from Figure 10, shown here for comparison. See Figure 10 for more explanation.

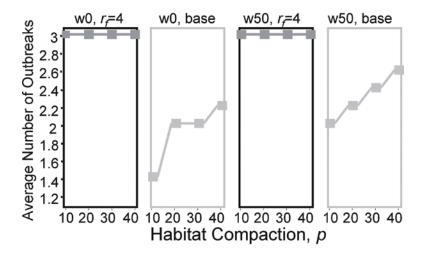


Figure 13: The effect of habitat compaction (p), patch size (w) and free flight  $(r_p)$  on the average number of outbreaks (op) in Experiment II. Note that outbreak frequency increases significantly with free flight  $(r_p)$  status (p<0.0001), but no other experimental effects are significant.

## 3.5 Experiment II extension - sensitivity without free flight

Overall, kill rate in the absence of free flight is very high (Experiment II). Does habitat compaction, communication distance or patch size have an effect if overall reproductive rate is not so high? To test the effect of decreased overall success on the relationship between success and habitat compaction in the case with no free flight we repeated Experiment II with a lower reproductive rate ( $r_w$ =3). We expect that when kill rate is not uniformly high, kill rate and outbreak frequency should depend on habitat compaction, communication distance, and patch size.

Decreasing winter reproductive rate  $(r_w)$  significantly alters the relationship between kill rate and habitat compaction, and this effect varies significantly with patch size  $(kr \sim r_w * w * p, n = 240, p < 0.0124)$  (Figure 14). Decreasing  $r_w$  also alters the relationship between kill rate, communication distance, patch size and response variable  $(g_w \text{ and } d_w)$   $(kr \sim r_w * d * w * rpt, n = 240, p < 0.0002)$  (Figure 14). Decreasing reproductive rate also decreased the frequency of outbreaks overall  $(op \sim r_w * rpt, n = 20, p < 0.0001)$  (Figure 15). Essentially, when both reproductive rate and patch size are small, kill rate and outbreak probability increase with both patch size and habitat compaction as expected (Figure 14 and 15). Increasing patch size largely eliminates these effects.

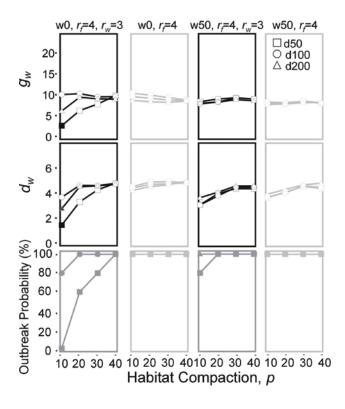


Figure 14: The effect of winter reproductive rate  $(r_w)$  on the results from Experiment II. The interaction between reproductive rate  $(r_w)$ , communication distance (d), patch size (w), and response variable  $(d_w, g_w)$  is significant (p=0.0002), as is the interaction between habitat compaction (p) and response variable  $(d_w, g_w)$  (p<0.0001) and the interaction between reproductive rate  $(r_w)$ , patch size (w) and habitat compaction (p) (p=0.0124). Symbols are filled as in Figure 10.

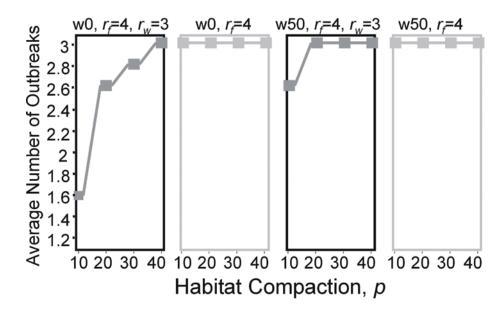


Figure 15: The effect of winter reproductive rate  $(r_w)$  on the results from Experiment II. Outbreak frequency increases significantly with free flight  $(r_p)$  status (p<0.0001), and no other experimental effects are significant.

#### 4.0 Discussion

#### 4.1. Beetle ecology

Before considering what the results of this study say about mountain pine beetles, it is worth asking whether model behaviour is reasonable. First, the striking result from Experiment II is that beetles with a free flight period outbreak less often and less quickly. In general, biologists are well advised to be wary of claims that behaviour should be advantageous just because it is done (Gould and Lewontin 1979). However, since the trait is common but not ubiquitous among beetle taxa, and variable within beetle populations (Borden et al 1986), it seems most prudent to assume free flight is advantageous under at least some circumstances, and to be wary of a model that predicts otherwise. One problem might be that we allow unreceptive beetles to die, but not land. Our free flight period is perhaps too long in some circumstances, and dispersal during the free flight period might not be random. In general, we presume that real free flight behaviour is more subtly formulated to help beetles avoid overcrowding while not making them lost. Free flight may also have other functions not addressed by the model.

Model behaviour is unrealistic in several other ways: real beetles can easily spread over 250 ha in 4 or 5 years, rather than 9 or 10; and real beetles often do not kill 100% of susceptible trees. Given the simplicity of the model, the number of uncertain parameters and relationships, and the limited amount of sensitivity analysis and calibration, we expect a model such as this to be more useful as a tool to explore our state of knowledge and assumptions, and to make general rather than precise predictions.

At face value, the prediction is that beetles are relatively insensitive to small-scale landscape heterogeneity because they have a free flight period. A more cautious conclusion is that the spatial dynamics of infestation spread at small scales can depend strongly on relatively minor details of dispersal behaviour, such as free flight. All beetle dispersal models omit a great deal of detail, and most have omitted free flight. Readers and modelers should be aware of this omission.

We began with a model developed by others, then considered the assumptions underlying that model, and the consequences for our particular question. Working from another model made us more acutely aware of our own assumptions, and the challenges to be overcome. We differ from the MPBpde on two main points other than the inclusion of a free flight period. First, we assume that beetles continue to attack trees in an area until all susceptible hosts are killed. In the MPBpde, there is no explicit link between the pheromone concentration at which an area becomes unattractive and the point at which landing rate declines to zero, so areas with very low resistance (and therefore low beetle capacity) might never become unattractive, while areas with high resistance might become unattractive before host resistance is overcome.

Our second concern was that a flying beetle in an area with no trees, ringed on all sides by areas full of nesting beetles will tend to remain in place, even though there is nowhere for it to land. In contrast, we assume the antiaggregants discourage beetles from landing, but not from dispersing through an area. Beetles also may fly above the canopy to avoid overcrowding (Safranyik 1978; Safranyik et al. 1989; Safranyik et al. 1992).

This work required that we clarify expectations about what scale of landscape heterogeneity might affect beetles. Since pheromones are only effective over 50 or 100 m, the possibility is that patchiness at quite small scales may be important. Although we have used the term "landscape", in forestry management the question is about patchiness within, rather than between, forest stands.

Ultimately, questions about the effect of pattern on beetles can only be answered by experimentation. Experimentally manipulating forest "landscapes" tens or hundreds of kilometers across is difficult. Forest stands, on the other hand, are quite manageable experimental units. Indeed, there have been a number of harvesting experiments that indicate thinned stands are less susceptible to mountain pine beetles (Cahill 1978; Cole et al. 1983; Mitchell et al. 1983; Waring and Pitman 1985; McGregor et al. 1987; Amman et al. 1988; Mitchell 1994). Thinning is one extreme in a range of forest harvest options. Exploring the

effects of a wider range of cutting patterns is both feasible and potentially interesting. Finally, note that the nature of the non-susceptible area probably makes a difference: pheromone plumes are more likely to be disrupted by turbulent air in open areas than under a forest canopy.

It would be theoretically interesting to see how the susceptibility of patchy stands differs from both thinned and unharvested stands. It might also be operationally useful. Even though thinning is known to reduce stand susceptibility, it is difficult and expensive, and thus not extensively done. Cutting larger patches may be more operationally feasible.

#### 4.2. Movement modeling

Others have recognized that organisms often deviate from random walk or correlated random walk movement patterns (Okubo 1980; Levin 1992; Turchin 1998; Okubo and Levin 2001), but the consequences of deviation have not been systematically studied. Real-world successes with correlated random walks and related diffusion approximations indicate that these models are robust to some behavioural complexity, and more general than their simplicity suggests (Levin 1992; Turchin 1998). However, the strong effect of free flight here indicates that not all behavioural complexity is without consequence. It would be interesting to better understand the limits of diffusion and correlated random walk models, and the types of behaviour that render these models inadequate.

Rather than testing arbitrary possibilities, it is probably more productive to focus on understanding the behaviour of particular organisms in particular situations. However, studies of the dispersal of particular organisms do not automatically contribute to more general understanding. Thus far, general insight from individual-based models has been hampered by lack of systematic reference to theory and lack of a common framework that would make it possible to meaningfully compare models (Turchin 1998; Grimm et al. 1999). In response, some authors have advocated a hierarchical approach, where understanding is sought by comparing the behaviour of simple models to those incrementally more complex (Turchin 1998; Grimm et al. 1999). We have attempted this approach, and agree that it is a fine ideal. We note, however, that many of the criticisms of individual-based modeling (and spatio-temporal simulation in general), while valid, are not easily addressed. Although many simulation efforts would benefit from more comprehensive experimentation and more systematic reference to the framework of classical theoretical ecology (Grimm 1999), such attempts should proceed with a realistic view of the technical and computational challenges. Spatial simulation is computationally demanding.

Apart from theoretical concerns and concerns over inadequate data, complex individual-based models have been criticized because they are hard to develop, hard to communicate, and hard to understand (Grimm et al. 1999). Individual-based models implemented with general purpose programming languages have been plagued by software bugs, awkward software design, and general incomprehensibility (Grimm et al. 1999). In response to these challenges, a number of software tools have been developed to help separate the details of model form from model implementation, and thus make implementation, verification and communication easier (Lorek and Sonnenschein 1998, 1999; Fall and Fall 2001). We did not try alternative methods for implementing this model, so cannot compare their relative merits. Anecdotally, the process of incremental component testing and experimentation was greatly aided by the SELES modeling tool (Fall and Fall 1999; Fall and Fall 2001). We draw particular attention to the possibility for discrete entities (individual beetles) and continuous quantities (pheromones) to spread and change at different time-scales in the same model, as this is the first such "mixed" model implemented in SELES to be published.

# **Acknowledgements**

Thanks to Terry Shore, John Borden, Carl Schwarz, Ian Bercovitz, Joe Fall, Marie-Josee Fortin, and Laurence Lee. This work was funded by a National Science Research Council Post-Graduate Scholarship, a Canadian Forest Service PGSA supplement, a Paul and Helen Trussell Scholarship, and a Simon Fraser University Applied Sciences Graduate Fellowship to J. Hughes. Assistance was also provided through a Global Forest Grant to K. Lertzman.

#### References

- Agresti, A. 1996. An Introduction to Categorical Data Analysis. John Wiley & Sons, New York.
- Amman, G.D. 1978. Biology, ecology and causes of outbreaks of the mountain pine beetle in lodgepole pine forests. Pages 39-53 *in* A. A. Berryman, G. D. Amman, and R. W. Stark, editors. Theory and Practice of Mountain Pine Beetle Management in Lodgepole Pine Forests. University of Idaho, Pullman, WA.
- Amman, G.D.; Logan J.A. 1998. Silvicultural control of mountain pine beetle: prescriptions and the influence of microclimate. American Entomologist 44(3): 166-177.
- Amman, G.D.; McGregor, M.D.; Dolph, R.E.Jr. 1984. Mountain pine beetle. USDA Forest Service, Washington, D.C., Forest Insect and Disease Leaflet 2
- Amman, G.D.; McGregor, M.D.; Schmitz, R.F.; Oakes, R.D. 1988. Susceptibility of lodgepole pine to infestation by mountain pine beetles following partial cutting of stands. Canadian Journal of Forest Research 18(6): 688-695.
- Amman, G.D.; Safranyik, L. 1985. Insects of lodgepole pine: impacts and control. Pages 99-105 *in* D.M. Baumgartner, R.G. Krebill, J.T. Arnott, and G.F. Weetman, editors. Lodgepole pine: the species and its management. Proceedings of a conference, May 8-10, 1984, Spokane, WA. Washington State University, Pullman.
- Andren, H. 1994. Effects of habitat fragmentation on birds and mammals in landscapes with different proportions of suitable habitat: a review. OIKOS 71: 355-366.
- Bender, D.J.; Contreras, T.A.; Fahrig, L. 1998. Habitat loss and population decline: a meta-analysis of the patch size effect. Ecology 79(2): 517-533.
- Bentz, B.J.; Logan, J.A.; Amman, G.D. 1991. Temperature dependent development of the mountain pine beetle (Coleoptera: Scolytidae) and simulation of its phenology. The Canadian Entomologist 123: 1083-1094.
- Berryman, A.A. 1978. A synoptic model of the lodgepole pine/mountain pine beetle interaction and its potential application in forest management. Pages 98-105 *in* A. A. Berryman, G. D. Amman, and R. W. Stark, editors. Theory and Practice of Mountain Pine Beetle Management in Lodgepole Pine Forests. University of Idaho, Pullman, WA.
- Biesinger, Z.; Powell, J.; Bentz, B.; Logan, J. 2000. Direct and indirect parameterization of a localized model for the mountain pine beetle: lodgepole pine system. Ecological Modelling 129(2-3): 273-296.
- Bolker, B.M. 1999. Analytic models for the patchy spread of plant diseases. Bulletin of Mathematical Biology 61: 849-874.
- Borden, J.H.; Hunt, D.W.A.; Miller, D.R.; Slessor, K.N. 1986. Orientation in forest Coleoptera: an uncertain outcome of responses by individual beetles to variable stimuli. Pages 97-109 *in* T.L. Payne, M.C. Birch, and C.E.J. Kennedy, editors. Mechanisms in Insect Olfaction. Clarendon Press, Oxford.
- Borden, J.H. 1993. Uncertain fate of spot infestations of the mountain pine beetle, *Dendroctonus pondero-sae* Hopkins. The Canadian Entomologist 125(1): 167-169.
- Borden, J.H.; Ryker, L.C.; Chong, L.J.; Pierce, H.D.; Johnston, B.D.Jr.; Oehlschlager, A.C. 1987. Response of the mountain pine beetle, *Dendroctonus ponderosae* Hopkins (Coleoptera: Scolytidae), to five semiochemicals in British Columbia lodgepole pine forests. Canadian Journal of Forest Research 17: 118-128.

- Byers, J.A. 1988. Upwind flight orientation to pheromone in western pine beetle tested with rotating windvane traps. Journal of Chemical Ecology 14(1): 189-198.
- Cahill, D.B. 1978. Cutting strategies as control measures of the mountain pine beetle in lodgepole pine in Colorado. Pages 188-191 *in* Kibbee, D.L., A.A. Berryman, G.D. Amman, and R.W. Stark, editors. Theory and Practice of Mountain Pine Beetle Management in Lodgepole Pine Forests. University of Idaho, Pullman, WA.
- Cantrell, R.S.; Cosner, C. 1991. The effects of spatial heterogeneity in population dynamics. Journal of Mathematical Biology 29: 315-338.
- Capasso, V. 1993. Mathematical structures of epidemic systems. Springer-Verlag, Berlin.
- Cappuccino, N.; Lavertu, D.; Bergeron, Y.; Regniere, J. 1998. Spruce budworm impact, abundance and parasitism rate in a patchy landscape. Oecologia 114(2): 236-242.
- Choudhury, J.H.; Kennedy, J.S. 1980. Light versus pheromone-bearing wind in the control of flight direction by bark beetles. Physiological Entomology 5: 207-214.
- Clark, W. C.; Jones, D.D.; Holling, C.S. 1979. Lessons for ecological policy design: a case study of ecosystem management. Ecological Modeling 7: 1-53.
- Cole, W.E.; Cahill, D.B.; Lessard, G.D. 1983. Harvesting strategies for management of mountain pine beetle infestations in lodgepole pine: preliminary evaluation, East Long Creek demonstration area, Shoshone National Forest, Wyoming. USDA Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT., Research Paper INT-318
- Cook, N.R.; Ware, J.H. 1983. Design and analysis methods for longitudinal research. Annual Review of Public Health 4: 1-23.
- Coulson, R.N.; McFadden, B.A.; Pulley, P.E.; Lovelady, C.N.; Fitzgerald, J.W.; Jack, S.B. 1999. Heterogeneity of forest landscapes and the distribution and abundance of the southern pine beetle. Forest Ecology and Management 114(2/3): 471-485.
- Fall, A.; Fall, J. 1999. SELES user and modeller documentation. Available online at: http://www.gowlland.ca/
- Fall, A.; Fall, J. 2001. A domain-specific language for models of landscape dynamics. Ecological Modelling 137: 1-21.
- Fall, J. 1998. Reconstructing the historical frequency of fire: a modelling approach to developing and testing methods. Masters Thesis in Resource Management. Simon Fraser University, Burnaby, BC.
- Fortin, M.J.; Dale, M. 2005. Spatial Analysis: A Guide For Ecologists. Cambridge University Press, Cambridge.
- Franklin, J.F.; Forman, R.T.T. 1987. Creating landscape patterns by forest cutting: ecological consequences and principles. Landscape Ecology 1(1): 5-18.
- Gara, R.I.; Geiszler, D.R.; Littke, W.R. 1984. Primary attraction of the mountain pine beetle to lodgepole pine in Oregon. Annals of the Entomological Society of America 77(4): 333-334.
- Geiszler, D.R.; Gallucci, V.F.; Gara, R.I. 1980. Modeling the dynamics of mountain pine beetle aggregation in a lodgepole pine stand. Oecologia 46: 244-253.
- Gould, S.J.; Lewontin, R.C. 1979. The spandrels of San Marco and the panglossian paradigm: a critique of the adaptionist programme. Proceedings of the Royal Society of London, Series B 205(1161): 581-598.

- Grimm, V. 1999. Ten years of individual-based modelling in ecology: what have we learned and what could we learn in the future? Ecological Modelling 115: 129-148.
- Grimm, V.; Wyszomirski, T.; Aikman, D.; Uchmanski, J. 1999. Individual-based modelling and ecological theory: sythesis of a workshop. Ecological Modelling 115: 275-282.
- Hastings, A. 1977. Spatial heterogeneity and the stability of predator-prey systems. Theoretical Population Biology 12: 37-48.
- Heibeler, D. 2000. Populations on fragmented landscapes with spatially structured heterogeneities: landscape generation and local dispersal. Ecology 81(6): 1629-1649.
- Helland, I.S.; Anderbrant, O.; Hoff, J.M. 1989. Modelling bark beetle flight: a review. Holarctic Ecology 12: 427-431.
- Helland, I.S.; Hoff, J.M.; Anderbrant, O. 1984. Attractions of bark beetles (Coleoptera: Scolytidae) to a pheromone trap. Journal of Chemical Ecology 10(5): 723-752.
- Keitt, T.H.; Urban, D.L.; Milne, B.T. 1997. Detecting critical scales in fragmented landscapes. Conservation Ecology [online] 1(1): 4. Available online at: http://www.consecol.org/vol1/iss1/art4.
- Kostyk, B.C.; Borden, J.H.; Gries, G. 1994. Photoisomerization of antiaggregation pheromone verbenone: biological and practical implications with respect to the mountain pine beetle, *Dendroctonus ponderosae* Hopkins (Coleoptera: Scolytidae). Journal of Chemical Ecology 19(8): 1749-1759.
- Levin, S.A. 1992. The problem of pattern and scale in ecology. Ecology 73(6): 1943-1967.
- Li, H.; Reynolds, J.F. 1994. A simulation experiment to quantify spatial heterogeneity in categorical maps. Ecology 75(8): 2446-2455.
- Logan, J.A.; White, P.; Bentz, B.J.; Powell, J. A. 1998. Model analysis of spatial patterns in mountain pine beetle outbreaks. Theoretical Population Biology 53: 236-255.
- Lorek, H.; Sonnenschein, M. 1998. Object-oriented support for modelling and simulation of individual-oriented ecological models. Ecological Modelling 108: 77-96.
- Lorek, H.; Sonnenschein, M. 1999. Modelling and simulation software to support individual-based ecological modeling. Ecological Modelling 115: 199-216.
- Ludwig, D.; Walker, B.; Holling, C.S. 1997. Sustainability, stability, and resilience. Conservation Ecology [online] 1(1): 7. Available online at: http://www.consecol.org/vol1/iss1/art7 (viewed Oct 4, 2006).
- McGregor, M.D.; Amman, G.D.; Schmitz, R.F.; Oakes, R.D. 1987. Partial cutting lodgepole pine stands to reduce losses to the mountain pine beetle. Canadian Journal of Forest Research 17: 1234-1239.
- Maclaughlan, L.E.; Brooks, J.E. (editors) 1994. Strategies and tactics for managing the mountain pine beetle, *Dendroctonus ponderosae*. British Columbia Forest Service, Kamloops Region Forest Health, Kamloops, B.C.
- Mawby, W.D.; Hain, F.P.; Doggett, C.A. 1989. Endemic and epidemic populations of southern pine beetle: implications of the two-phase model for forest managers. Forest Science 35(4): 1075-1087.
- May, R.M. 1978. Host-parasitoid systems in patchy environments: a phenomenological model. Journal of Animal Ecology 47: 833-843.
- Mitchell, J.L. 1994. Commercial thinning of mature lodgepole pine to reduce susceptibility to mountain pine beetle. FRDA Report 224, Forest Engineering Research Institute of Canada, Vancouver, B.C.

- Mitchell, R.G.; Preisler, H.K. 1991. Analysis of spatial patterns of lodgepole pine attacked by outbreak populations of the mountain pine beetle. Forest Science 37(5): 1390-1408.
- Mitchell, R.G.; Waring, R.H.; Pitman, G.B. 1983. Thinning lodgepole pine increases tree vigor and resistance to mountain pine beetle. Forest Science 29(1): 204-211.
- Okubo, A. 1980. Diffusion and ecological problems: mathematical models. Springer-Verlag, Berlin.
- Okubo, A. and Levin, S. 2001. Diffusion and ecological problems: modern perspectives. 2nd Ed. Springer, New York.
- Polymenopoulos, A.D.; Long, G. 1990. Estimation and evaluation methods for population growth models with spatial diffusion: dynamics of mountain pine beetle. Ecological Modelling 51: 97-121.
- Potvin, C.; Lechowicz, M.J.; Tardif, S. 1990. The statistical analysis of ecophysiological response curves obtained from experiments involving repeated measures. Ecology 71(4): 1389-1400.
- Powell, J.; Kennedy, B.; White, P.; Bentz, B.; Logan, J.; Roberts, D. 2000. Mathematical elements of attack risk analysis for mountain pine beetles. Journal of Theoretical Biology 204(4): 601-620.
- Powell, J.; Tams, J.; Bentz, B.; Logan, J. 1999. Theoretical analysis of "switching" in a localized model for mountain pine beetle mass attack. Journal of Theoretical Biology 194(1): 49-63.
- Powell, J.A.; Logan, J.A.; Bentz, B.J. 1996. Local projections for a global model of mountain pine beetle attacks. Journal of Theoretical Biology 179(3): 243-260.
- Powell, J.A.; McMillen, T.; White, P. 1998. Connecting a chemotactic model for mass attack to a rapid integro-difference emulation strategy. SIAM Journal of Applied Mathematics 59(3): 547-572.
- Raffa, K.F.; Berryman, A.A. 1986. A mechanistic computer model of mountain pine beetle populations interacting with lodgepole pine stands and its implications for forest managers. Forest Science 32(3): 789-805.
- Ratkowsky, D.A. 1990. Handbook of nonlinear regression models. M. Decker, New York, NY.
- Reeve, J.D. 1988. Environmental variability, migration, and persistence in host-parasitoid systems. The American Naturalist 132(6): 810-836.
- Rodriguez, D.J.; Torres-Sorando, L. 2001. Models of infectious diseases in spatially heterogeneous environments. Bulletin of Mathematical Biology 63: 547-571.
- Safranyik, L. 1978. Effects of climate on weather on mountain pine beetle populations. Pages 77-84 *in* Kibbee, D.L., A.A. Berryman, G.D. Amman, and R.W. Stark, editors. Theory and Practice of Mountain Pine Beetle Management in Lodgepole Pine Forests. University of Idaho, Pullman, WA.
- Safranyik, L. 1989. Mountain pine beetle: biology overview. USDA Forest Service, Intermountain Research Station, Ogden, UT., General Technical Report.
- Safranyik, L. 1999. A population dynamics model for the mountain pine beetle, *Dendroctonus ponderosae* Hopk. (Coleoptera: Scolytidae). Natural Resources Canada, Canadian Forest Service, Pacific Forestry Centre, Victoria, B.C., Information Report BC-X-386.
- Safranyik, L.; Linton, D.A.; Silversides, R.; McMullen, L.H. 1992. Dispersal of released mountain pine beetles under the canopy of a mature lodgepole pine stand. Journal of Applied Entomology 113: 441-450.
- Safranyik, L.; Linton, D.A. 1993. Relationships between catches in flight and emergence traps of the mountain pine beetle, *Dendroctonus ponderosae* Hopk. (Col.: Scolytidae). Journal of the Entomological Society of British Columbia 90: 53-61.

- Safranyik, L.; Silversides, R.; McMullen, L.H.; Linton, D.A. 1989. An empirical approach to modeling the local dispersal of the mountain pine beetle *Dendroctonus ponderosae* Hopk. in relation to sources of attraction. Journal of Applied Entomology 108(5): 498-511.
- Samman, S.; Logan, J. 2000. Assessment and response to bark beetle outbreaks in the rocky mountain area. USDA Forest Service, Rocky Mountain Research Station, Fort Collins, CO., General Technical Report RMRS-GTR-62.
- Saunders, D.A.; Hobbs, R.J.; Margules, C.R. 1991. Biological consequences of ecosystem fragmentation: a review. Conservation Biology 5(1): 18-33.
- Sewell, G. 1988. The Numerical Solution of Ordinary and Partial Differential Equations. Academic Press, Inc. San Diego, CA.
- Shore, T. L.; Safranyik, L. 1992. Susceptibility and risk rating systems for the mountain pine beetle in lodgepole pine stands. Forestry Canada, Pacific Forestry Centre, Victoria, B.C., Information Report BC-X-336. 12 p.
- Shore, T.L.; Safranyik, L.; Lindgren, B.S. 1992. The response of mountain pine beetle (*Dendroctonus ponderosae*) to lodgepole pine trees baited with verbenone and exo-brevicomin. Journal of Chemical Ecology 18(4): 533-541.
- Shrimpton, D.M.; Thomson, A.J. 1983. Growth characteristics of the lodgepole pine associated with the start of mountain pine beetle outbreaks. Canadian Journal of Forest Research 13(1): 137-144.
- Simberloff, D. 1988. The contribution of population and community biology to conservation science. Annual Review of Ecology and Systematics 19: 473-511.
- Smith, A.T.; Peacock, M.M. 1990. Conspecific attraction and the determination of metapopulation colonization rates. Conservation Biology 4(3): 320-323.
- Taylor, A.D. 1990. Metapopulations, dispersal, and predator-prey dynamics: an overview. Ecology 71(2): 429-433.
- Thompson, W.A.; Vertinsky, I.B.; Wellington, W.G. 1981. Intervening in pest outbreaks: simulation studies with the western tent caterpillar. Researches on Population Ecology 23(1): 27-38.
- Thomson, A.J. 1987. Comparison of lodgepole pine yield tables. Canadian Journal of Forest Research 17(9): 1110-1114.
- Thomson, A.J.; Safranyik, L.; Shrimpton, D.M.; Whitney, H.S. 1984. A theory of mountain pine beetle population response to weather-induced changes in host resistance. Pages 128-135 *in* L. Safranyik, editor. The role of the host in the population dynamics of forest insects. USDA Forest Service and Environment Canada, Canadian Forestry Service, Pacific Forest Research Centre, Victoria, B.C.
- Thomson, A.J.; Shrimpton, D.M. 1984. Weather associated with the start of mountain pine beetle outbreaks. Canadian Journal of Forest Research 14(2): 255-258.
- Tischendorf, L.; Fahrig, L. 2000. On the usage and measurement of landscape connectivity. Oikos 90(1): 7-19.
- Turchin, P. 1989. Population consequences of aggregative movement. Journal of Animal Ecology 58: 75-100.
- Turchin, P. 1991. Translating foraging movements in heterogeneous environments into the spatial distribution of foragers. Ecology 72(4): 1253-1266.

- Turchin, P. 1998. Quantitative analysis of movement: measuring and modeling population redistribution in animals and plants. Sinauer Associates, Inc., Sunderland, MA.
- Turchin, P.; Thoeny, W.T. 1993. Quantifying dispersal of southern pine beetles with mark-recapture experiments and a diffusion model. Ecological Applications 3(1): 187-198.
- Turner, M.G.; Gardner, R.H.; Dale, V.H.; O'Neill, R.V. 1989. Predicting the spread of disturbance across heterogeneous landscapes. Oikos 55: 121-129.
- Turner, M.G.; Romme, W.H. 1994. Landscape dynamics in crown fire ecosystems. Landscape Ecology 9(1): 59-77.
- Turner, M.G.; Romme, W.H.; Gardner, R.H. 1999. Prefire heterogeneity, fire severity, and early postfire plant reestablishment in subalpine forests of Yellowstone National Park, Wyoming. International Journal of Wildland Fire 9(1): 21-36.
- Turner, M.G.; Gardner, R.H.; O'Neill, R.V. 2001. Landscape ecology in theory and practice: pattern and process. Springer, New York.
- von Ende, C.N. 2001. Repeated measures analysis: growth and other time-dependent measures. Pages 134-157 *in* S.M. Scheiner and J. Gurevitch, editors. Design and analysis of ecological experiments. Oxford University Press, New York.
- Waring, R.H.; Pitman, G.B. 1985. Modifying lodgepole pine stands to change susceptibility to mountain pine beetle attack. Ecology 66(3): 889-897.
- White, P.; Powell, J. 1997. Phase transition from environmental to dynamic determinism in mountain pine beetle attack. Bulletin of Mathematical Biology 59(4): 609-643.
- White, P.; Powell, J. 1998. Spatial invasion of pine beetles into lodgepole forests: a numerical approach. SIAM Journal of Scientific Computing 20(1): 164-184.
- Whitehead, R.; Martin, P.; Powelson, A. 2001. Reducing stand and landscape susceptibility to mountain pine beetle. British Columbia Ministry of Forests, Forest Practices Branch, Victoria, B.C. Available online at: http://www.for.gov.bc.ca/hfp/pubs/standman/MPB\_SDM.pdf.
- Wood, D.L.; Stark, R.W.; Waters, W.E.; Bedard, W.D.; Cobb, J. 1985. Treatment tactics and strategies. in W.E. Waters, R.W. Stark, and D.L. Wood, editors. Integrated Pest Management in Pine-Bark Beetle Ecosystems. John Wiley and Sons, New York.
- Zollner, P.A.; Lima, S.L. 1999. Search strategies for landscape level interpatch movements. Ecology 80(3): 1019-103

# **Tables**

Table 1a: Definitions of terms from landscape ecology.

Landscape ecology te	rms
landscape	Area that is spatially heterogeneous in at least one factor of interest (Turner et al. 2001).
cell	In this study, land is represented as matrix (raster) of square cells. Each cell characterized by a spatial location (row and column) and the spatial resolution (50 by 50 m). Landscape attributes (e.g. suitability of habitat for beetles or the number of beetles at a given time) are spatially homogeneous within cells.
scale	The spatial or temporal dimension of an object or process, characterized by both grain (or resolution) and extent (Turner et al. 2001). For example, the (spatial) extent of our landscapes is 5 km by 5 km, and the resolution is 50 m by 50 m. The (temporal) extent of our simulations is 20 years, and the temporal resolution of the mountain pine beetle dynamics differs from the temporal resolution of the pheromone dynamics.
habitat	Cells that contain at least some pine that is susceptible to beetles are classified as habitat.
patch	In general, surface area that differs from its surroundings in nature or appearance, or, on a gridded landscape, a contiguous group of cells of the same mapped category (Turner et al. 2001). In this paper, the mapped category of interest is suitability for beetles (habitat or non-habitat). We use an 8-neighbour rule, so diagonally adjacent cells are considered to be touching one another.
spatial heterogeneity/ patchiness	In general, complexity and variability of a system property in space (Li and Reynolds 1994). System properties that may vary on our binary (habitat/non-habitat), gridded landscapes include the size and shape of patches. The most homogenous possible landscape is one where habitat cells are distributed uniformly across the landscape, and a landscape where cells are distributed randomly also has low heterogeneity. Heterogeneity increases as cells are grouped together into patches of varying size and complexity. We also informally refer to spatial heterogeneity in this context as patchiness.
spatial autocorrelation	If a variable is spatially autocorrelated then the relationship among the values of a given variable is a function of the spatial distances between them or their locations in space. Hence, the notion of spatial dependence implies that there is a lack of independence among data from nearby locations. Spatial autocorrelation is estimated by comparing the value of a variable at one location with those at given distances apart (termed spatial lag or distance interval) (Fortin and Dale 2005). Positive spatial autocorrelation at short lag distances indicates that similar values are grouped together in space.
fragmentation	The breaking apart of habitat into smaller patches, or patches separated by greater distance.
cluster	A cluster consists of one or more patches that are less than some threshold distance apart from one another (Keitt et al. 1997). In this case, the threshold distance of interest is the distance across which beetles can effectively communicate (communication distance).

Table 1b: Definitions of terms from population ecology and terms specific to this paper.

Population ecology terms	
infestation/ outbreak	Mountain pine beetle population dynamics are characterized by endemic periods, when beetles persist in low numbers and kill few trees, interspersed with periods of rapid population increase, when beetles kill many trees. Loosely, we refer to a rapid rise in beetle population and damage in one area as an <i>infestation</i> . If local infestations arise in many areas independently or spread affect larger areas, this landscape-level phenomenon is an <i>outbreak</i> .
aggregation/ congregation	Turchin (1998) defines aggregation as population redistribution that leads to an uneven spatial distribution of organisms so that some spatial localities are characterized by elevated population densities (aggregations) and others by decreased density. In contrast, congregation is aggregation as a result of behavioural responses of organisms to conspecifics. Thus, to <i>congregate</i> means to gather <i>together</i> ; as opposed to <i>aggregate</i> , which is to gather <i>at</i> some locality. Congregating organisms may respond to neighbours using visual, acoustic, or chemical (pheromones) stimuli, or indirectly to population density cues, such as feeding damage on a host plant. According to this definition, mountain pine beetles congregate. However, aggregation is the term commonly used in mountain pine beetle literature, so we use the term aggregative to describe beetle dispersal.
Terms specific to this pap	per
communication distance	Communication distance is the distance across which pheromones effectively spread and beetles effectively communicate (Section 1.4).
habitat compaction	Habitat compaction increases as the area across which (a fixed amount of) habitat is dispersed decreases (Section 2.2.1). As habitat compaction increases the distance between patches decreases, given constant patch size and shape.
centre cluster size	Centre cluster size is the area of habitat within the cluster at the centre of each landscape (Section 3.2). In simulation runs, infestations originate in the centre of the landscape. If beetle spread is strictly limited by whether or not beetles can communicate across gaps then the final infestation extent should be equal to the centre cluster size in each landscape.
free flight	Free flight is a period of flight before beetles become receptive to pheromones (Section 1.3.3).
nesting beetle	Nesting beetles are beetles that have committed to staying at a location. In reality, once beetles have landed, they may decide to resume flying if a tree seems unsuitable. For simplicity, we only consider a beetle to be nesting once it has both landed and decided to remain at a location (Section 3.1.2, 3.1.5).

Table 2: Probabilities for transitions between five beetle modes. Each beetle can be either waiting to emerge, flying and unreceptive to pheromones or kairomones (uflying), flying and receptive to pheromones and kairomones (rflying), nesting once they have landed, or dead, and each beetle may undergo one mode transition per beetle time-step  $\Delta t$ . pf is the length of the annual flight period (in hours),  $\omega_j$  is the death rate of flying beetles (per hour) and  $r_f$  is the rate at which unreceptive free flying beetles become receptive to pheromones (per hour). The landing rate of flying beetles at each location (landing rate, depends on the number of nesting beetles (N), host capacity (K) and host resistance (R) at that location. See text for more detailed explanation.

#### **Beetle Mode at Time t**

	Booto modo at 1mo t							
		waiting	uflying	rflying	nesting	dead		
	waiting	1-∆t/p <sub>f</sub>	∆t/p <sub>f</sub>	0	0	0		
; t-∆t	uflying	0	1- $\Delta t \omega_1 - (1-\Delta t \omega_1)\Delta t r_f$	$(1-\Delta t \omega_1) \Delta t r_f$	0	$\Delta t \omega_1$		
Time	rflying	0	0	1- $\Delta t \omega_1$ – (1- $\Delta t \omega_1$ ) $\Delta t$ (landing rate <sub>i,j</sub> )	(1- $\Delta t \ \omega_{_1}$ ) $\Delta t \ (landing \ rate_{_{i,j}})$	$\Delta t \omega_1$		
	nesting	0	0	0	1	0		
	dead	0	0	0	0	1		

Table 3: Model parameters and their associated base values. Units are:  $\mu g = 10$ -6 grams; ha = hectare; fh = flight-hour; kMPB = thousands of mountain pine beetles.

Parameters	rameters Description		Units	Source
Landscape				
$P_I$	initial beetle population	1250	kMPB	arbitrary
$R_{0}$	average initial resistance	0.5		arbitrary
$R_{dev}$	standard deviation of initial resistance	0.3		arbitrary
$K_0$	average initial carrying capacity	1000	kMPB	Biesinger et al. 2000
$K_{dev}$	maximum deviation of initial carrying capacity from average initial carrying capacity	500	kMPB	arbitrary
M	spatial extent	100	cells	reasoning
Н	spatial resolution	0.25	ha cell <sup>-1</sup>	reasoning
Time				
$p_f$	length of annual flight period	40	fh year <sup>-1</sup>	Logan et al. 1998
$\Delta t$	beetle time step	0.25	fh	reasoning
$\Delta t_a$	pheromone time step	1/500	fh	reasoning
Flight				
$\alpha_m$	approximately average turning angle in the absence of pheromones	35°	degrees	arbitrary
$a_l$	pheromone sensitivity parameter	1.3		arbitrary
$a_r$	minimum detectable pheromone concentration	0.1	μg ha <sup>-1</sup>	Geiszler et al. 1980
$a_s$	pheromone sensitivity parameter	4		arbitrary
$r_f$	free flight parameter	0.65	fh <sup>-1</sup>	Borden et al. 1986
Landing				
$r_m$	maximum possible landing rate	0.8	fh <sup>-1</sup>	arbitrary
$r_b$	maximum pioneer landing rate (landing = $r_b$ when $R = 0$ and $N = 0$ )	0.2	fh <sup>-1</sup>	arbitrary
$r_{I}$	controls the rate at which landing increases with N	20		arbitrary

Table 3 continued on the following page.

Table 3 continued:

Landing continued				
$r_2$	controls the effect of R on pioneer landing rate	2		arbitrary
$r_3$	determines the persistence of preference for low R areas as N increases	0.3		arbitrary
$r_5$	controls the rate at which landing decrease with crowding	4		arbitrary
Pheromone				
$a_m$	maximum possible pheromone production rate	20	μg ha <sup>-1</sup>	Biesinger et al. 2000
$a_l$	controls the rate at which pheromone production declines with N	20		arbitrary
$a_3$	determines the effect of R on the N value at which pheromone production reaches half of maximum	0.3		arbitrary
$a_6$	ensures that beetles that land in areas with no resistance $(R=0)$ do not produce pheromones, but beetles that land anywhere else produce pheromones at maximum rate until resistance is overcome.	0.001 (very small)		reasoning
$b_a$	pheromone diffusivity	0.685	ha fh <sup>-1</sup>	Biesinger et al. 2000
$\delta_a$	pheromone decay rate	180	μg ha <sup>-1</sup>	Biesinger et al. 2000
Survival				
$\omega_l$	flying beetle death rate	0.01	fh <sup>-1</sup>	Biesinger et al. 2000
$r_w$	maximum per capita beetle reproduction rate, referred to as winter reproductive rate $(F_{waiting, n}/N_{n-1})$	3.5	year <sup>-1</sup>	arbitrary
$s_I$	controls the rate at which beetle success increases with N	$a_{I}$		reasoning
\$3	determines the effect of R on the N value at which beetle success reaches half of maximum	$a_3$		reasoning
$S_4$	determines the average reproductive rate at carrying capacity (N = K)	$r_5 + 1$		reasoning
<b>S</b> 5	determines the rate at which reproductive success decreases to minimum as N approaches K.	$r_5$		reasoning
S <sub>6</sub>	ensures beetles landing in areas with no resistance reproduce at maximum, but beetles landing elsewhere are not successful unless resistance is overcome.	$a_6$		reasoning

Table 4. Summary of experimental results. Abbreviations: Experimental factors are denoted by w (patch size), p (habitat compaction), and d (communication distance). kr indicates overall kill rate response, rpt refers to response variable (gw or dw – if rpt is significant then gw and dw differ in their response to the experimental factors), and op is overall outbreak probability. Other parameters are as in Table 3. n: Two sample sizes given for each experiment are the number of model runs (between\*within subject factors) and, in brackets, the number of landscape instances (between subject factors only). Design: All experiments are have fully crossed factorial designs. For example, "kr ~ dlwlp" indicates that all independent effects (d, w, p) and all possible interactions  $(d^*w, d^*p, w^*p, d^*w^*p)$  were tested for. Test Results: Only significant interactions are shown. Statistical p values are distinguished from habitat compaction references by italics, bolding, and context. Summary of Effects: Up (1), down (1) or sideways (↔) arrows indicate positive, negative or neutral relationships, respectively, between the response variable (right side) and the experimental factor (left side). Notably strong and consistent effects are shown in bold (1). Effects that are conditional upon the value of other factors are qualified with the key word given, followed by a list of conditions. For qualitative variables, the keyword when followed by one or more conditions indicates that response in these conditions is higher (1) or lower (1) than in other conditions.

Experiment	n	Design	Test Results	Summary of Effects	Notable Results
I	240	kr ~	$kr \sim d*rpt, p < 0.0001$	kr <b>↑</b> d	• no effect of patch size (w)
base case	(80)	d w p rpt	$kr \sim p, p = 0.0003$	kr↑ p	• effect of patch compaction (p)
Section 3.3,					not as strong as expected.
Figures 9 and 10		op ~ w   p	no op effects significant		
II	240	kr ∼	$kr \sim r_f^* d^* rpt, \ p < 0.0001$	when $r_f$ = base, effects are as in	• free flight (rj=base)
the effect of free	(40)	r <sub>f</sub>   d w p rpt	$kr \sim r_f * p* rpt, \ p = 0.0486$	Experiment I.	disadvantageous to beetles unless
flight			$kr \sim r_f * d* w, p = 0.0053$	$kr \uparrow when r_f = 4 given d=0-50$	communication distance (d) is
				$kr \uparrow when r_f = 4$	high, or communication distance i
				given d=50-100 and p<30	low and patch compaction is also
				$kr \downarrow when r_f = 4$	low.
				given d=50-100 and p>20	effect of communication distance
Section 3.4,		op ~		$kr \downarrow when r_f = 4$	(d) and patch compaction (p) not
Figures 11 and 12		r <sub>f</sub>   w   p	op $\sim r_f$ , $p < 0.0001$	given d=100-200	robust to absence of free flight
8		11.11.15	op 14, <i>p</i> 10.0001	op $\uparrow$ when $r_f = 4$	$(r_{j}=4).$
II extension	240	kr ∼	$kr \sim r_w^* d^* w^* rpt, \ p = 0.0002$	kr ↑ p given r <sub>w</sub> =3, w0, d=0-50	• patch compaction (p), patch size
sensitivity without	(40)	r <sub>w</sub>  d w p rpt	$kr \sim p^* rpt, \ p < 0.0001$	$kr \uparrow p $ given $r_w$ =3, w0, d=50-100	(w) and communication distance
free flight			$kr \sim r_w^* w^* p$ , $p = 0.0124$		(d) effects significant in the
Section 3.5,		op ~	_ <del>-</del>		absence of free flight (r <sub>j</sub> =4) when
Figures 13 and 14		r <sub>w</sub> l w l p	op $\sim r_{\rm w}, \ p < 0.0001$		overall success rate is reduced.

43

## **Appendix**

This appendix contains further information about the form of model equations.

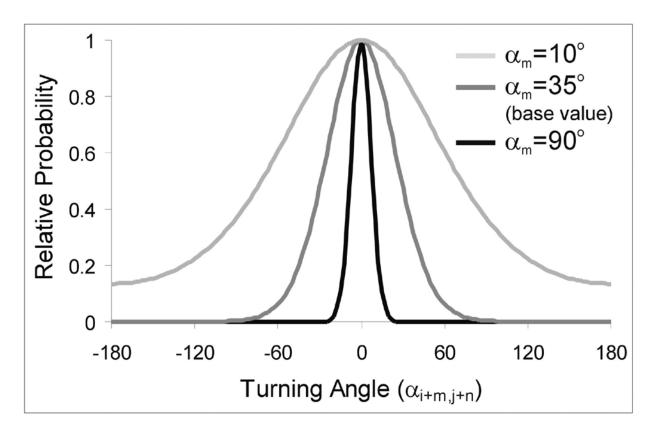


Figure A.1: Effect of  $\alpha_m$  on the distribution of mountain pine beetle turning angles in free flight or the absence of chemical attractants (Equation 1). The quantity  $(1-\cos(a_{i+m,j+n}))$  is exponentially distributed with parameter  $\alpha_m$ , which is approximately equal to the average turning angle when  $\alpha_m$  is small (~<60°). Average turning angle does not continue to increase indefinitely with  $\alpha_m$  because the finite range of possible turning angles truncates the exponential distribution.

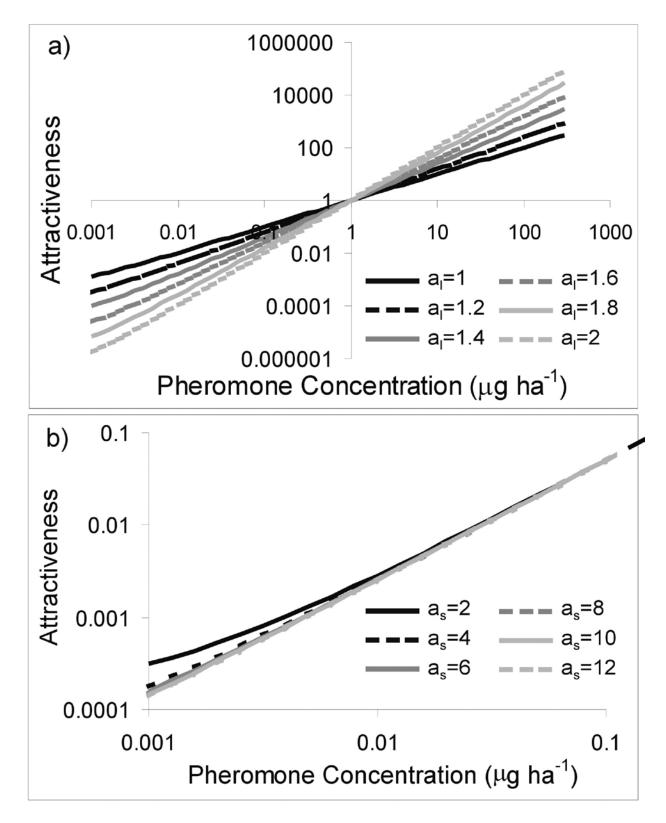


Figure A.2: The sensitivity of attractiveness to the gradient sensitivity parameters,  $a_l$  (a), and  $a_s$  (b) (Equation 3). The attractiveness parameter,  $a_l$ , controls the sensitivity of mountain pine beetles to pheromone gradients, and  $a_s$  determines the relative preference of beetles for low pheromone over no pheromone areas.

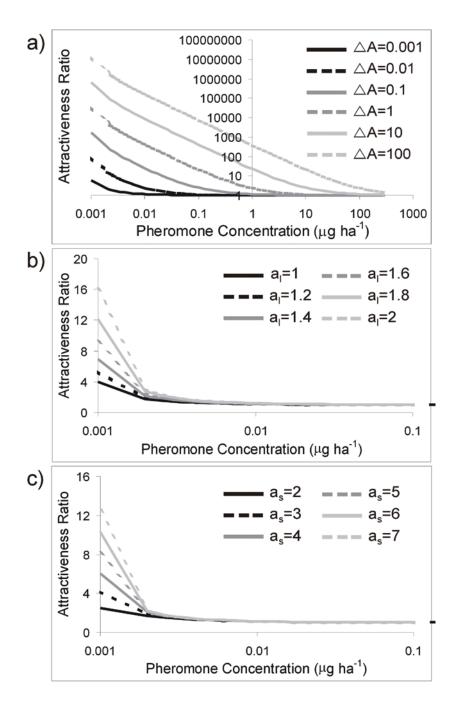


Figure A.3: Sensitivity of the relative attractiveness of pheromones to gradient size ( $\Delta A$ ) and gradient sensitivity parameters  $a_i$  and  $a_s$  (Equations 2 and 3). The probability a beetle will choose high pheromone location over a lower pheromone location is given by the attractiveness ratio, ( $Attractiveness_{high}$ /  $Attractiveness_{low}$ ). The attractiveness ratio depends on (a) the difference in pheromone concentration between the two locations ( $\Delta A$ ), (b,c) the attractiveness parameters  $a_i$  and  $a_s$ , and the base pheromone concentration (x-axis). The attractiveness function is designed so that beetles are more sensitive to differences in pheromone concentration when pheromone concentrations are low.

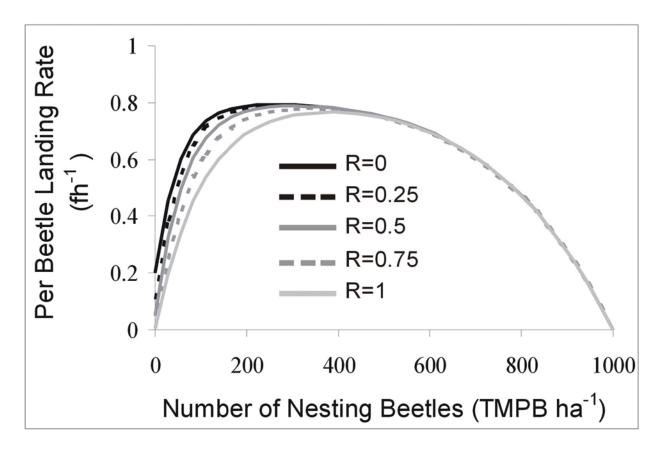


Figure A.4: Sensitivity of the relationship between per beetle landing rate and the number of nesting beetles (N) to host resistance (R) (Equations 4 and 5). Carrying capacity K is set to 1000 kMPB ha<sup>-1</sup> in this example.

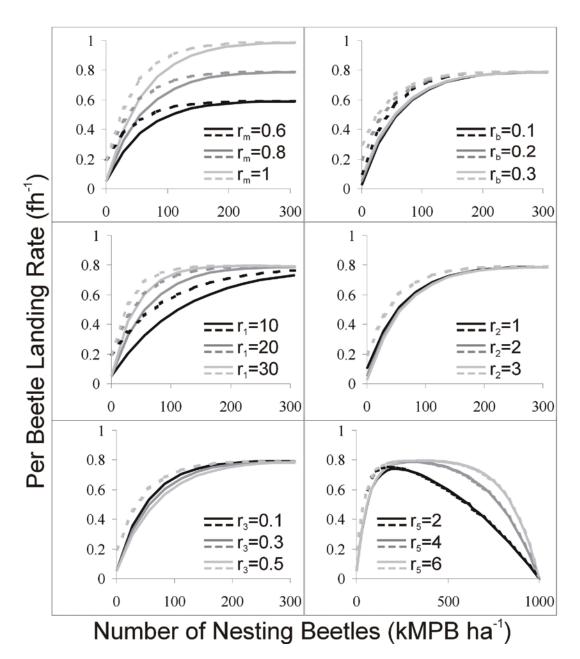


Figure A.5: Sensitivity of landing rate to six model parameters (Equations 4 and 5). Solid lines indicate the case where resistance (R) is 0.5. Dotted lines indicate R = 0.  $r_b$  is the maximum possible landing rate of pioneer beetles,  $r_m$  is the maximum possible landing rate,  $r_l$  determines how fast landing rate increases with the number of nesting beetles,  $r_2$  controls the preference of beetles for low resistance areas,  $r_3$  determines how the relative preference of flying beetles from low resistance areas persists as the number of nesting beetles increases, and  $r_5$  determines the sensitivity of flying beetles to crowding.

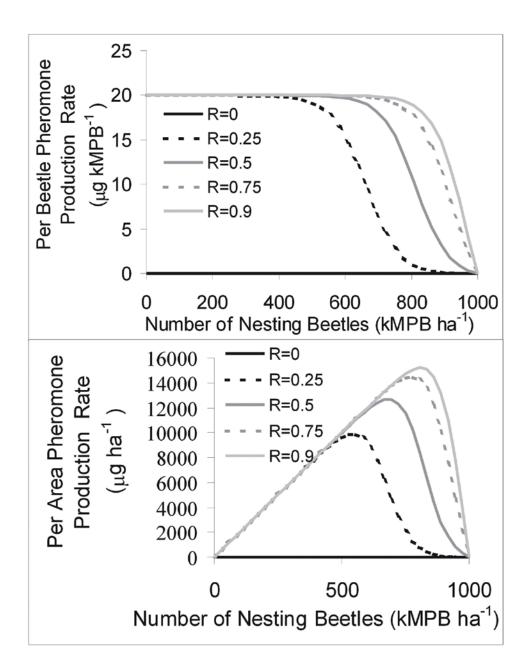


Figure A.6: Sensitivity of per beetle and per area pheromone production rates to host resistance (R) (Equation 6).

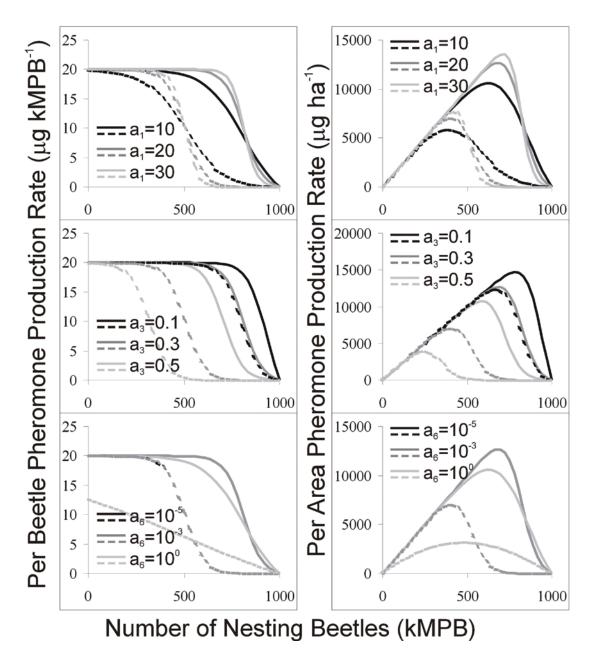


Figure A.7: Sensitivity of per capita and per area pheromone production to three model parameters (Equation 6). Solid lines indicate the case where resistance (R) is 0.5. Dotted lines indicate R = 0.  $a_1$  determines the steepness of production decline,  $a_3$  determines the rate of increase, and  $a_6$  ensures that initial per beetle pheromone production is maximum unless host resistance is very near zero.

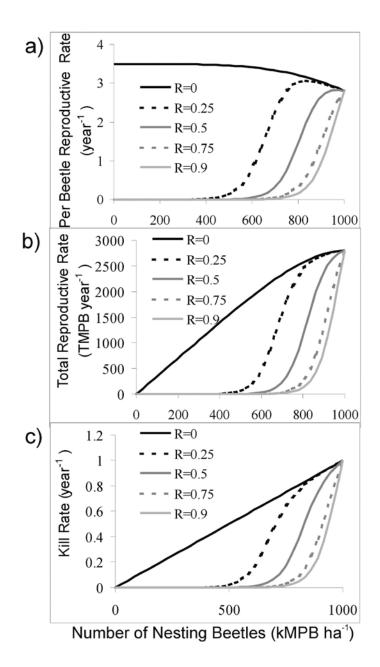


Figure A.8: Sensitivity of (a) per capita reproductive rate, (b) total reproductive rate, and (c) forest kill rate to host resistance (R) (Equations 10-12).

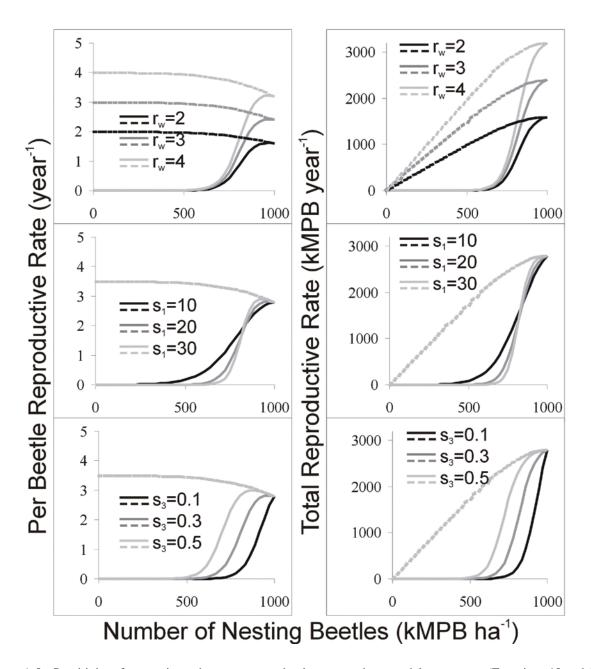


Figure A.9: Sensitivity of per capita and per area reproductive rate to three model parameters (Equations 10 and 11).  $r_w$  determines the maximum per capita reproductive rate in the absence of resistance or competition, and  $s_1$  and  $s_3$  together determine how strongly beetle success is impacted by host resistance. Solid lines indicate the case where resistance (R) is 0.5. Dotted lines indicate R = 0.

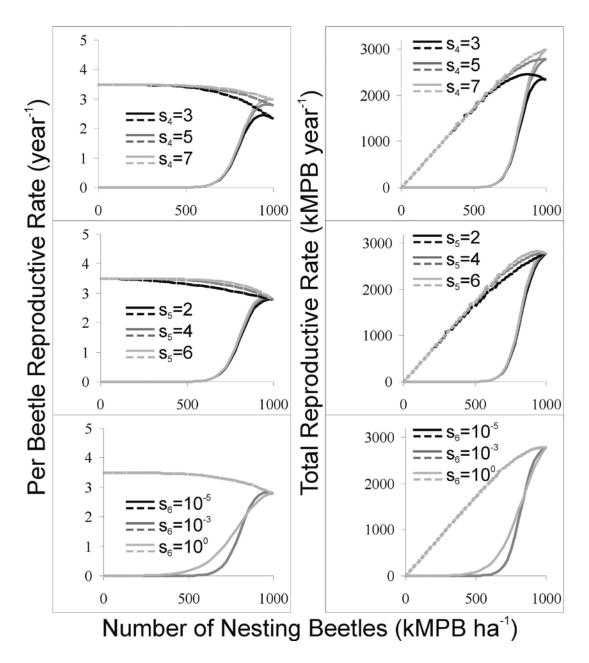


Figure A.10: Sensitivity of per capita and per area reproductive rate to 3 model parameters (Equations 10 and 11).  $s_4$  determines the reproductive rate of beetles at carrying capacity,  $s_4$  determines the rate at which reproductive success decreases due to crowding, and  $s_6$  ensures that beetles are not successful unless there are enough beetles to overcome host resistance. Solid lines indicate the case where resistance (R) is 0.5. Dotted lines indicate R = 0.

### **Canadian Forest Service Contacts**

For more information about the Canadian Forest Service, visit our website at www.nrcan.gc.ca/cfs-scf/ or contact any of the following Canadian Forest Service establishments

1 Atlantic Forestry Centre P.O. Box 4000 Fredericton, NB E3B 5P7 Tel.: (506) 452-3500 Fax: (506) 452-3525

atl.cfs.nrcan.gc.ca/

Atlantic Forestry Centre – District Office Sir Wilfred Grenfell College Forestry Centre University Drive

Corner Brook, Newfoundland A2H 6P9
Tel.: (709) 637-4900 Fax: (709) 637-4910

2 Laurentian Forestry Centre 1055 rue du P.E.P.S., P.O. Box 3800 Sainte-Foy, PQ G1V 4C7 Tel.: (418) 648-5788 Fax: (418) 648-5849 www.cfl.scf.rncan.gc.ca/ Great Lakes Forestry Centre
P.O. Box 490 1219 Queen St. East
Sault Ste. Marie, ON P6A 5M7
Tel.: (705) 949-9461 Fax: (705) 759-5700
www.glfc.cfs.nrcan.gc.ca/

A Northern Forestry Centre 5320-122nd Street Edmonton, AB T6H 3S5
Tel.: (403) 435-7210 Fax: (403) 435-7359 nofc.cfs.nrcan.gc.ca/

Pacific Forestry Centre
506 West Burnside Road
Victoria, BC V8Z 1M5
Tel.: (250) 363-0600 Fax: (250) 363-0775
www.pfc.cfs.nrcan.gc.ca/

6 Headquarters 580 Booth St., 8th Fl. Ottawa, ON K1A 0E4 Tel.: (613) 947-7341

Fax: (613) 947-7396



To order publications on-line, visit the Canadian Forest Service Bookstore at: bookstore.cfs.nrcan.gc.ca