# **Boreal Shield Forest Disturbance and Recovery Trends using** Landsat Time Series

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### 2 Highlights:

- We develop and apply a Landsat time series methodology for spectral recovery
- East & West Ca. Boreal Shield section's disturbance/recovery trends are
   compared
  - Disturbance and recovery trends are different in each Boreal Shield sections
  - Differences in recovery are apparent from spatial tabulation of recovery rates
  - Selection of spectral index affects the characterization of recovery through time
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# 11 Abstract:

- 12 Monitoring forest recovery following disturbance is important for both forest
- 13 management as well as assessing possible climate change impacts on forest dynamics.
- 14 To do so, an improved understanding of forest recovery processes and their
- 15 relationship to remotely sensed spectral measures of recovery is required. Our objective
- 16 in this research is to develop and apply a methodology for using Landsat time series to
- 17 characterize forest recovery using spectral recovery trajectories. We focus our efforts in
- 18 the Canadian Boreal Shield ecozone where a known geographic east to west distinction
- in disturbance regimes remains to be quantified. Results show that forest recovery
- 20 following a stand replacing disturbance is detectable and quantifiable using a dense
- 21 Landsat time series of spectral reflectance. All Tasseled Cap indices were found to
- 22 capture an element of forest recovery following disturbance, with Wetness offering
- additional information on increasing vegetation structure and complexity. Tasseled Cap
- component trajectories of recovery show clear differences in both disturbance detection
- and forest recovery across the east and west Boreal Shield sections. The Cohen's d
  similarity metric indicated large differences (d > .08) in Wetness and Greenness-based
- spectral recovery trajectories when comparing the two Boreal Shield sections with the
- 28 East Boreal Shield having markedly more above average recovery (+2 std. dev. from the
- 29 mean) than the West. Based on our spectral recovery results, we also observe that forest
- 30 recovery varies over the entire ecozone and is different between the east and west
- 31 Boreal Shield forests

#### 32 Introduction:

- 33 While variable based upon definitions, Canadian boreal forests represent about 30% of
- 34 global boreal forests (Brandt et al. 2013, NRCan 2014). Wildfire is the primary
- disturbance agent in boreal forests, though insect infestation, windthrow, harvest, ice
- 36 and snow related damage are all additional factors that can also occur with varying
- 37 spatial pervasiveness and severity (Johnson 1996, Payette 1992, Brandt et al. 2013).
- 38 Long-term averages indicate that wildfire occurs over approximately 2,000,000 hectares
- 39 per year in Canada (Stocks et al 2002) with the most common being stand replacing
- 40 crown and severe surface fires (Heinselman 1981). This fire regime has led to boreal tree
- species being well adapted to these fire disturbances, and often depending on fire to
- 42 release the growing space from constraints imposed by the overstory (Oliver and
- 43 Larson 1996, Brandt 2009, Bonan and Shugart 1989). As a consequence the boreal forest
- is a patchwork of forest structures, types, and ages, which is critical to maintaining the
- diversity and sustainability of the Canadian boreal zone (Bonan and Shugart 1989,
- 46 Heinselman 1981).
- 47
- 48 The Canadian boreal can be stratified into ecozones which provide a spatial framework
- 49 of ecologically homogenous units delineating distinct areas based on biophysical factors
- 50 (Ecological Stratification Working Group 1996, Wiken 1986). The Boreal Shield ecozone
- is the largest unit stretching from Newfoundland into northern Saskatchewan with
- 52 much of the area inaccessible and in a wilderness state. Despite the reported
- 53 homogeneity of this ecozone, studies have suggested that the ecozone could be split
- along an east / west divide based on a number of distinct factors. Kurz et al (1992)
- suggested the ecozone be split due to differing ecoclimatic conditions. Kull et al (2006)
- split the ecozone citing colder and drier climes in the western section than the east. To
- 57 characterize fire in the Canadian boreal forest, the ecozone was divided into eastern and
- 58 western sections by Stocks et al (2002). Likewise, citing differing climatic patterns and
- forest processes, Andrew et al (2012) divided the Boreal Shield into two separate
- 60 sections to aid their analysis of protected areas in Canada's boreal zone.
- 61

As a result of differing levels of accessibility (Andrew et al 2012), large areas of the

- 63 Canadian Boreal Shield are not subject to direct anthropogenic influences with up to
- 64 35% and 40% of the forested East and West Boreal Shield ecozone sections not subject to
- any forest management practices (Shvidenko & Apps 2006, Wulder et al 2007). Thus the
- 66 Boreal Shield can be further divided into two management zones: a generally southern
- 67 managed portion that enjoys a relative wealth of descriptive data, and a generally
- 68 northern unmanaged section that is less well characterized. The different management
- cones create a divide in forest information, which leaves a large part of the ecozone
- 70 uncharacterized. The lack of spatially explicit forest information for more than a third of

- the ecozone is a problem when assessing both current and future climate change (Price
- 72 et al 2013). Increased productivity in temperature limited predominantly eastern boreal
- 73 areas (Boisvenue and Running 2006), and increased quantity and severity of wildfire
- 74 disturbances in drier western boreal areas (Flannigan et al 2005) is expected,
- 75 highlighting the need for disturbance mapping and monitoring of the subsequent
- 76 return of vegetation.
- 77

78 Remote sensing technology has demonstrated the capacity to monitor large areas and

- can offer marked insights into how different regions are potentially changing, both in
- 80 terms of pressures on disturbance regimes as well as the subsequent recovery (Powers
- et al, 2015, Beck and Goetz 2011, Berner et al 2011, Xu et al 2013, Pickell et al 2014). To
- 82 date, emphasis has been placed on mapping forest disturbance using remote sensing
- technologies (Eidenshink et al 2007, Guindon et al 2014, Steyaert et al 1997). Although
- 84 forest disturbances are well characterized by remote sensing (Wulder et al 2009,
- 85 Potapov et al 2008, Kasischke et al 2011, Loboda et al 2012), explicit monitoring of post
- 86 disturbance forest conditions for signals of recovery using a remotely sensed time series
- of imagery is a relatively recent phenomenon (Griffiths et al 2014, Kennedy et al 2012,
- 88 Schroeder et al 2011, Chen et al 2011, Gitas et al 2012). Forest recovery has typically
- **89** been defined as the reestablishment of forest biomass, or canopy structure, following
- 90 disturbance (Oliver and Larson 1996), making it a process rather than a state, and
- 91 therefore observable using time series of remote sensing data (Kennedy et al 2012,
- 92 Hermosilla et al 2015, Huang et al 2010, Zhu et al 2012). It is increasingly understood
- **93** that characterization of the spectral recovery as observed from remote sensing of a
- 94 recently disturbed forest can aid in identifying and monitoring the progression of the
- **95** stand towards successful reestablishment and maturation (Schroeder et al 2011).
- 96 Interestingly, Frolking et al (2009) identified the need for remote sensing approaches to
- 97 examine the characterization of forest recovery to provide forest managers with
- **98** valuable information such as successional state (Shatford et al 2007) as well as
- 99 informing on regional or temporally varying recovery trends.
- 100

Forest recovery has been monitored using image based time series using a variety ofapproaches. The most common, used with varying success, is to develop a time series of

- 103 a given spectral index which is then related to biophysical parameters at a range of
- spatial scales. (Chu et al 2013). Griffiths et al (2014) assessed variability in forest
- recovery both spatially and temporally across political jurisdictions in the Carpathian
- 106 by defining spectral recovery as a combination of pre-disturbance spectral index values
- and the spectral magnitude of the disturbance. Kennedy et al (2012) defined spectral
- 108 recovery as a ratio of the magnitude of the disturbance event to spectral conditions five
- 109 years post disturbance and illustrated recovery differences between public and
- 110 privately owned lands in the Pacific Northwest of the United States. In the same study

- 111 Kennedy et al (2012) showed that drier areas experienced slower recovery as opposed
- more moist areas across management regimes, ownership, and political lines. Thus
- 113 meaningful information about forest recovery across large areas can be derived from
- 114 remote sensing imagery based time series, providing useful data and facilitating
- 115 decisions about forest recovery by managers and researchers alike.
- 116

117 Landsat time series analysis offers an effective approach to monitor large forested areas

- 118 (Hermosilla et al 2015, Huang et al 2010, Zhu et al 2012) like the Boreal Shield ecozone
- for disturbances and can enable characterization of the subsequent vegetative recovery(Kennedy et al 2010). In this research we compare and contrast spectral forest recovery
- 121 trajectories following forest disturbance across the east and west sections of the Boreal
- 122 Shield ecozone by developing and then applying a methodology using Landsat time
- series imagery. To meet this objective we examine stand replacing disturbances and the
- 124 subsequent spectral signals of recovery by constructing spectral trajectories of recovery.
- 125 These trajectories are then compared to nominally undisturbed forested signals, and the
- 126 extent and recovery rate trends examined and compared across the ecozone.
- 127

# 128 Study Area:

- 129 The Boreal Shield Ecozone (Figure 1) is generally characterized by rolling and hilly
- topography with many small lakes, streams and rocky outcrops. A precipitation
- 131 gradient exists with higher amounts (1000mm) in the coastal east and less (400mm) in
- the more continental west (Urquizo et al 2000, Ecological Stratification Working Group
- 133 1996). July average maximum temperatures are 13°C for both sections, however the east
- typically has less severe winters with January average minimum temperatures of -1°C
- compared to the west with -20°C (Urquizo et al 2000, Ecological Stratification Working
- **136** Group 1996). The Boreal Shield ecozone is dominated by forests of black (*Picea mariana*)
- and white spruce (*Picea glauca*) with the more southerly portions having a wider mix of
- broadleaf treed vegetation such as white birch (*Betula papyrifera*), trembling aspen
- 139 (*Populus tremuloides*) as well as an array of needle leaf species such as white (*Pinus*
- 140 *strobus*), and red (*Pinus resinosa*) and jack pine (*Pinus banksiana*) (Ecological Stratification
- 141 Working Group 1996). As discussed, fire and harvest are the primary agents of stand
- replacing forest disturbances (Bergeron et al 2004, Bergeron et al 2000), with fires
- 143 occurring more often over larger areas in the west than the east (Stocks et al 2002). Less
- 144 common are disturbances caused by insect infestations, wind and storm related
- 145 damage, and disease (Urquizo et al 2000).For the purpose of this study we follow the
- 146 division of the Boreal Shield ecozone established by Kull et al (2006), who cited the
- 147 colder and drier climate in the west resulting in differing forest processes. As depicted148 in Figure 1 the ecozone can be divided by southern managed zone, where commercial

- forestry activities are present, and a northern largely unmanaged (non-commercial)zone.
- 150 Z 151
- 152 Methods
- 153 Data

### 154 Landsat

155 The Landsat World Referencing System (WRS-2) acquisition grid (185 x 185km) was

- used to create a stratified random non overlapping sample of Landsat scene footprints
- 157 within the Boreal Shield ecozone (Tucker et al 2000, Wulder et al 2001, Healey et al 2006,
- 158 Kennedy et al 2010, Masek et al. 2013). Scenes were required to be at least 50% within
- the Boreal Shield ecozone, with portions outside of the ecozone clipped from further
- analysis. The Earth Observation for Sustainable Development of Forests (EOSD) landcover dataset was used to mask water bodies and to define areas of forest cover within
- the selected Landsat scenes (Wulder et al 2008). Deciduous, coniferous, and mixed
- 163 forested type classes from the EOSD were collapsed into one category to create a binary
- 164 forest layer, which was then used as a second filter to ensure that scenes were
- 165 dominated by forest cover (> 80%). The scenes were further filtered by forest
- 166 management zone, removing any scene that was not at least 50% within one
- 167 management zone or another; if a scene covered both management zones, it was
- 168 clipped to the portion within the largest management zone. Scenes were then randomly
- selected from the final candidate pool, while minimizing differences in total area as
- possible. Out of a possible 185 scenes covering the Boreal Shield ecozone, a total of 8
- 171 Landsat scenes were selected covering over 200,000 km<sup>2</sup> or 25 % of the ecozone's
- 172 forested area (Table 1).
- 173

## 174 Image Processing

- 175 In summary, each Landsat scene was subject to radiometric correction, cloud masking,
- 176 yearly compositing, and then temporal segmentation using the LandTrendr algorithm
- 177 (Kennedy et al. 2012). Temporal trajectory metrics were then extracted from the
- temporal segmentation results and used to characterize recovery following stand
- 179 replacing disturbances and compared between the east and west sections.
- 180 181

# Image Selection, Preprocessing, and Compositing

- 182 Once selected, the imagery was extracted from the Landsat archive based on two
- 183 criteria: (i) an acquisition date within the typical boreal zone growing season (which
- 184 corresponds to Julian days 152-243 and the calendar months of June, July, and August);
- (ii) cloud cover < 75%. A total of 1220 scenes were selected from the archive. The
- distribution of imagery over the years for each scene is displayed in Figure 2. All
- imagery was ortho-rectified and converted to surface reflectance using the Landsat

- Ecosystem Disturbance Adaptive Process (LEDAPS) (Masek et al 2006). A cloud mask 188
- was produced for each image using the Function of mask (Fmask) algorithm (Zhu and 189
- Woodcock 2012). Pixels delineated as cloud, cloud shadow, ice and snow were masked 190
- and not used in the remainder of the analysis. 191
- 192
- An annual time series for each of the eight Landsat scenes was constructed from the 193
- multiple images per year using rules developed by Kennedy et al (2010): i) for each year 194
- a base image was selected based on its proximity to July 19th (Julian day 200), 195
- coinciding with the mid-growing season; ii) all other images in order of their proximity 196 to July 19th within the same year were then used to fill in any areas that had been 197
- masked in the base image. This compositing process was repeated for each of the 29 198 199 years in the analysis period.
- 200

- **Spectral Indices and Recovery**
- Recent studies have highlighted differences in spectral recovery depending on the 202 spectral index examined (Pickell et al 2015, Banskota et al 2014, Cuevas-Gonzalez et al 203 2009, Epting and Verblya 2005). Frequently used in Landsat time series studies the 204 Tasseled Cap transformation is a linear transformation of Landsat's spectral bands into
- 205 three orthogonal axes of Brightness, Greenness, and Wetness using published 206
- coefficients (Crist and Ciccone 1984). Previous studies have found that Tasseled Cap 207
- Brightness provides an indication of overall pixel albedo; Greenness provides an 208
- indication of vegetation photosynthetic condition; and Wetness is sensitive to changes 209
- in forest structure (Cohen et al 1995, Collins & Woodcock 1996, Franklin et al 2000, 210
- Franklin et al 2002, Skakun et al 2003, Healey et al 2006). 211
- 212
- 213 After a stand replacing disturbance, the Tasseled Cap Components will have differing recovery trajectories responding to different properties of forest stands as the recovery 214 process progresses: 215
- Brightness relates to the general albedo (Crist and Ciccone 1984) of the 216
- land surface and values are typically high after a stand replacing 217
- disturbance and remain high due to the lack of absorption from 218
- vegetation cover and increased exposure of soil and non photsynthetic 219
- vegetation (Banskota et al 2014). Throughout forest stand recovery 220
- Brightness values will reduce with time due to increasing vegetation 221
- cover and subsequent increased absorption of visible and NIR 222
- wavelengths. This reduction in Brightness is typically observed as the 223
- stand reinitiates and growing space becomes increasingly occupied. As a 224
- result Brightness is uniquely suited to detect if a disturbed forest stand is 225
- 226 recovering and if vegetation is re-establishing post disturbance.

- Greenness contrasts the visible with the near and shortwave infrared 228 229 wavelengths. The absorption of visible radiation by vegetation, and scattering driven by leaf by structure in the near infrared combine to 230 produce high Greenness values as dense leafy forest canopies develop 231 (Crist and Ciccone 1984). Low Greenness values are typically due to lack 232 of photosynthetic vegetation, or vegetation that is under considerable 233 stress. After a stand replacing disturbance, Greenness would be 234 anticipated to increase quickly due to its sensitivity to green vegetation 235 (Pickell et al 2015). 236
- 237

Wetness was named for its response to moisture content in vegetation
and soils, and relies heavily upon the shortwave infrared bands. High
values of Wetness indicate an abundance of moist plant matter, while
lower values indicate reduced vegetation (Crist and Ciccone 1984).

- 242 Changes in forest structure have been related to changes in Wetness
- 243 (Cohen et al 1995, Cohen and Spies 1992, Hansen et al 2001), similar to
- the changes a forest stand undergoes during the recovery process. As a
- 245 forest stand recovers, structure becomes more complex, creating a
- 246 denser, moister, and multifaceted canopy with crown shadows,
- 247 branching, all of which are reflected by increased Wetness values. Thus
- 248 Wetness is well suited to observe forest recovery due to its ability to
- 249 detect structural changes over time.
- 250

Throughout the recovery process forest stands undergo many physical changes, which
can be observed using Landsat based Tasseled Cap time series. However, the capability
to detect change is limited over time by the asymptotical nature of reflected radiation
from a closed and dense forest canopy, as passive optical remote sensors can mostly
detect changes on the topmost reflective surface.

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# **Temporal Segmentation**

The LandTrendr temporal segmentation algorithm was used to reduce year to year 258 noise over the time series and segment the Wetness spectral values into distinct phases 259 of disturbance, recovery, and no change (Kennedy et al 2010, Powell et al 2010, Frazier 260 et al 2014). LandTrendr divides each pixel's time series using an iterative linear fitting 261 model, establishing breakpoints that define periods of no change, losses, and gains 262 within a temporal trajectory. Temporal trajectory metrics were then calculated from the 263 resultant temporal segments (Kennedy et al 2010, Frazier et al 2014, Pflugmacher et al 264 265 2012). To detect and isolate stand replacing disturbance events within the time series, 266 we followed Kennedy et al (2012) and applied thresholds to the largest (>700 change

- 267 magnitude) and most rapid (<3 years) changes in Wetness to identify severe and short</li>268 duration fire and harvest stand replacing disturbances.
- 269
- 270 Once disturbance events were identified using the Wetness trajectory, recovery
- trajectories that follow a disturbance event were isolated and only selected for further
- study if they exhibited a positive slope from the disturbance event to the end of the time
- 273 period, to ensure second disturbances did not occur. Using the year of disturbance, all
- recovery trajectories were then normalized by the year of the disturbance (see
- 275 Schroeder et al. 2011). The mean and standard deviation of the three tasseled cap
- components was computed annually for all recovery trajectories in the east and west
- 277 sections. For comparison, a stable, no change, forest mean threshold and standard
- deviation were also calculated from 18,652,249 no change pixels in the eastern section
- and 19,443,793 pixels in the western section.
- 280

281 Finally, each recovery pixel in each year was classified using the annual rate of recovery into one of five Wetness recovery classes: far above average recovery, above average 282 recovery, average recovery, below average recovery, and far below average recovery. 283 284 The far above average recovery represented values 2+ standard deviations above the section mean while above average values were between 1 - 2 standard deviations above 285 the average. The average recovery category was bounded by the ±1 standard deviation 286 from the recovery mean. Below average recovery was bounded by 1-2 standard 287 deviations below the average. Pixels experiencing recovery that was more than 2 288 standard deviations below the mean were classified as far below average recovery. The

- standard deviations below the mean were classified as far below average recovery. The
  amount of area for each disturbance and recovery category was also calculated as a
  percent of the total forest area under observation.
- 292

293

### Cohen's d statistic

294 Traditional statistical significance tests using very large sample sizes can be misleading, as any small difference would be significant at some level and often unimportant in 295 terms of causation (Johnson 1999). We required a statistic that did not take into account 296 297 our sample size and could determine meaningful levels of difference between our stable 298 forest signals and recovery trajectory, and then between both East and West Boreal Shield recovery mean sets. The Cohen's *d* statistic (Cohen 1988) can indicate similarity 299 300 or difference in datasets regardless of sample sizes. Cohen's *d* is a standardized statistic similar to z-scores, allowing the similarity or difference between two datasets to be 301 compared. Although not often used in remote sensing applications, Cohen's *d* has been 302 applied in the biological and ecological fields to determine likeness or difference 303 between samples. Studies have shown Cohen's *d* ability to determine differences 304 between datasets, when traditional significance testing may have shown less than 305 optimal results (Nakagawa et al 2007, Cavin et al 2013, Wesner et al 2012). 306

To calculate a *d* value, the means of two separate groups are subtracted and then divided by the pooled standard deviation as shown in the equation:

$$d = \frac{\bar{x}_1 - \bar{x}_2}{\sigma_{pooled}}$$

310 where  $\bar{x}_1$  and  $\bar{x}_2$  are the means of two groups, and  $\sigma_{pooled}$  is the pooled standard

deviation and *d* is a Cohen's *d* value. As described by Cohen (1988) and demonstrated

in further studies (Cavin et al 2013, Nakagawa et al 2007), predefined difference

313 thresholds set at d values of 0.2, 0.5, and 0.8 indicate small, moderate, or large

differences between two datasets respectively. We utilized Cohen's *d* to compare

recovery trends across the Boreal Shield ecozone and to establish the degree of

difference between stable and recovering vegetation, and difference between recovering

317 vegetation in each Boreal Shield section.

318

## 319 **Results**:

## 320 Wetness Recovery Trajectories and Cohen's *d* Values

After a stand replacing disturbance, Wetness has been shown to increase with time, and
often corresponds to increases in forest structure(Schroeder et al 2010). A positive
Wetness recovery trajectory occurs in both the eastern and western Boreal Shield

sections (Figure 3a). The eastern Boreal Shield initial Wetness value post disturbance is

far below the stable range of no change vegetation, but after 15 years of recoveryWetness values have increased to be within the stable range. Wetness values then

327 continue to increase until the end of the Landsat record when they resemble the stable,

no change vegetation mean. The western trajectory shows a similar trend, reaching the

stable range by year 16, however not exceeding the stable mean during the analysisperiod.

331

332 The eastern and western Cohen's *d* Wetness recovery trajectories (Figure 3b) indicate

that both sections are initially very different, but end the analysis period with relatively

- 334 little or no difference from their stable means. The *d* statistic for the eastern trajectory
- (2.09) indicates a large difference from its stable mean. After 25 years however the *d*
- value recovers to indicate no difference. The western trajectory's initial *d* value of 1.3

also indicates a large difference from its stable mean, but again after 22 years the

trajectory crosses into the zone of no change forest. The trajectory ends at a *d* value of

- **339** 0.37 indicating that the differences in Wetness of recovering vegetation and stable
- 340 vegetation is small.
- 341

342 When comparing the eastern and western trajectories to each other (Figure 3b) the

343 Cohen's *d* trajectory has an initial value of 0.01 indicating no initial difference between

- the two trajectories immediately following disturbance. After 20 years the recovery 344
- trajectories are markedly different from each other, and the comparison *d* trajectory 345
- ends at a value of 0.90 indicative of a large difference between the two Boreal Shield 346 sections.
- 347
- 348
- A summary of the Wetness, Greenness, Brightness, and associated Cohen's *d* trajectories 349
- are shown in Figures 3b-6b and Table 2. Each trajectory's start, midpoint and end 350
- difference levels and *d* values are presented for the eastern and western recovery 351
- trajectories, as well as the comparison *d* trajectory. 352
- 353 354

# Greenness Recovery Trajectories and Cohen's d Values

Like the wetness trajectories Greenness recovery is expected to increase as vegetation 355 returns to the stand, and photosynthetic capability of the stand increases. Both East and 356 West Boreal Shield Greenness trajectories conform to the expected positive trend 357 following disturbance (Figure 4a). The eastern trajectory begins below the stable mean 358 and within the stable range, but then quickly increases above the stable mean in the first 359 5 years. By year 19 recovery Greenness means are above the stable range, with the 360 trajectory stable at that level until the end of the time series. The western Greenness 361 trajectory has a very similar trend also starting below its stable range and increasing to 362 363 higher than stable mean values after 20 years of recovery, but not exceeding the stable

- range by the end of the trajectory. 364

365 Cohen's d Greenness recovery trajectories (Figure 4b) show similar shapes for the 366 eastern and western Boreal Shield, but have different *d* ranges indicating similar 367

recovery rates but different relationships to their no change stable means. The eastern *d* 368

- trajectory's initial value of 0.49 indicates little difference from its stable mean, but after 5 369
- years is not different and after 18 years is within the negative large difference range, 370
- where the trajectory ends with a *d* value of -1.04 indicating the recovery mean has 371
- 372 exceeded the stable mean. The western *d* trajectory's initial *d* value of 0.90 indicates a
- large difference from its stable mean, decreasing by year 17 to be within the no 373
- difference zone, remaining there until the end of the time series with a *d* value of 0.03. 374
- 375 This suggests that the Greenness signal from recovering vegetation is indistinguishable
- from stable mature forest vegetation after 18 years in the western Boreal Shield, 376
- however the eastern recovery trajectory shows more complicated trends. 377
- 378
- The eastern and western Greenness Cohen's *d* comparison trajectory initial value of 0.50 379
- indicates moderate difference (Figure 4b) between the two Greenness recovery 380
- trajectories. The *d* values quickly increase and by year 5 are within the large difference 381
- zone, remaining there until the end at a *d* value of 1.31. The high comparison *d* values 382

(>0.8) from 4 years post disturbance until the end of the comparison trajectory indicatethat Greenness recovery is largely different between both Boreal Shield sections.

385 386

## Brightness Recovery Trajectories and Cohen's dValues

As a stand recovers from a stand replacing disturbance Brightness will likely decrease 387 representing an increase in canopy cover and a reduction in overall albedo. The eastern 388 Brightness trajectory conforms to this expectation, however the western trajectory 389 shows no trend (Figure 5a). The eastern trajectory starts above its stable range and ends 390 above the stable mean. The western trajectory starts within the stable range and remains 391 392 there throughout its entire trajectory finishing at similar initial Brightness values. 393 Differences between the eastern and western trajectories are mirrored in their Cohen's d trajectories (Figure 5b). The eastern *d* trajectory's initial value of -0.84 indicates a large 394 395 difference from the stable mean. The trajectory then crosses into the moderate difference zone after 13 years and ends within the little difference zone with a *d* value of -0.49. The 396 western *d* trajectory's initial value of -0.36 is within the little difference zone and stays 397 398 within that zone for the remainder of the analysis period with a final *d* value of -0.48. 399

400 The Brightness comparison Cohen's *d* trajectory decreases over time with an initial

401 value of 0.51 within the moderate difference zone (Figure 5b). After 20 years the

402 trajectory crosses into the little difference zone and ends with a *d* value of 0.21 near the

403 lower bounds of the same zone signalling that recovering vegetation is always

somewhat different between the two sections, however becoming more similar as timesince disturbance time increases.

406

#### 407 408

### Areal Disturbance and Recovery Trends

The total yearly area disturbed in the eastern Boreal Shield sample shows a constant pattern of disturbances between 1985-2000 (Figure 6). Between 2001-2012 there is a marked increase in disturbed forest area. The western Boreal Shield shows a variable pattern of disturbance from 1985-2000 when compared to the eastern section, and again shows an increase in annual disturbance from 2001-2012.

414

The breakdown by trend rate for eastern and western Boreal Shield sections are shown

in Figures 8 and 9 respectively as a percent of total forest area. A relatively stable

disturbance trend in the eastern Boreal Shield from 1985 to 2000 is apparent from the

418 yearly totals (Figure 7). The increasing annual percentage of disturbance is also evident

from 2001-2012. The cumulative and classified recovery categories shown in Figure 7

420 demonstrates the variability in recovery rate after stand replacing disturbances. Most

421 apparent is the marked relative increases in the areas that are classified as far above

- 422 average recovery rate, defined as being 2 or more standard deviations above mean
- 423 Wetness recovery.
- 424
- 425 Variable disturbance trends characterize the western Boreal Shield between 1985-2000
- 426 as is evident from the yearly totals (Figure 8). An increasing amount of disturbance
- 427 from 2001-2012 is evident as well as an increase in the area undergoing far above
- 428 average recovery, similar to the eastern section. Throughout the same period, the
- amount of area classified as having an average rate of recovery remains stable from
- 430 2001-2012, at a level of about 0.5% of total forested area of the Boreal Shield.
- 431

# 432 Discussion:

- 433 In this research we examined the utility of Landsat time series analysis to characterize
- disturbance and recovery trends across the boreal forests of the Canadian Boreal Shield
- ecozone. We discuss the results under two broad headings: temporal trajectory
- 436 considerations and recovery characterization, and then our case study partitioning of
- 437 the East and West Boreal Shield.
- 438 439

# Temporal Trajectory Considerations & Recovery Characterization

- 440 Forest recovery refers to the reestablishment of key forest biophysical variables 441 following a disturbance event, and is a process not a state (Freilking et al 2009). Forest
- following a disturbance event, and is a process not a state (Frolking et al 2009). Forestrecovery rates can differ due to severity, frequency and type of disturbance, site
- 443 characteristics, and climate (Frolking et al 2009, Oliver and Larson 1996, Bolton et al
- 2015), Post disturbance forest recovery can take multiple pathways including (i) no
- 445 recovery (e.g. land use conversion), (ii) recovery towards a predefined stocking or
- structure (e.g. plantation forestry), or (iii) natural recovery. Critical to any study offorest disturbance and recovery, regardless of location, is the rate that forest vegetation
- is returning from the disturbance event. However, the multiple ways in which forest
- 449 recovery can be defined can lead to confusion when linking recovery to Landsat time
- 450 series data. Chief among the concerns is that ecological and silvicultural definitions of
- 451 forest recovery are not dependent on the actual presence of a mature forest, but more so
- 452 predicated on provision of ecosystem services or emergent characteristics regarding
- 453 treed vegetation height and canopy cover requirements.
- 454
- 455 Landsat time series can be used to characterize forest disturbance and forest recovery as
- 456 shown in this work. The information generated from the Landsat time series aids in
- 457 better understanding forest processes. Recovery extent, timing, and magnitude can now
- 458 be reported and compared across large areas to inform on regional or temporally
- 459 varying trends.
- 460

461 The terminology used by Oliver and Larson (1996) when describing forest stand

- 462 dynamics is particularly useful when describing spectral forest recovery trajectories, as
- it grounds spectral forest recovery in easier to understand terms that are more
- 464 generalizable and easier understood by a wider audience than spectral indices.
- 465 Oliver and Larson (1996) detail a generally applicable model of forest stand dynamics
- briefly summarized here with additional insight on how each Tasseled Cap component
- 467 is affected. A stand replacing disturbance opens growing space for new individuals468 (Oliver and Larson 1996) causing an increase in albedo that in turn is reflected by
- 469 increased in Brightness values. Individuals compete for resources until the canopy is
- 470 closed, which causes large increases in Greenness and moderate increases in Wetness
- 471 values, and a sharp decrease in Brightness. The stand reduces in density over time and
- the canopy increases in height (Oliver and Larson 1996) causing further increases in
- 473 Wetness over time. An understory redevelops as growing space and resources are
- released by the canopy (Oliver and Larson 1996), which is mostly undetectable to each
- 475 Tasseled Cap Index since it is occurring below the canopy. Shade tolerant individuals
- 476 grow from the understory into the canopy supplanting dominant canopy species
- 477 (Oliver and Larson 1996) potentially causing further change all three Tasseled Cap
- 478 indices. The stand is then governed by gap dynamics caused by small disturbances
- 479 (Oliver and Larson 1996 until a stand replacing disturbance occurs, which will restart480 the recovery process (Oliver and Larson 1996).
- While the spectral recovery trajectories described above are simplified archetypes, the 481 reality will be different as the combined elements that drive recovery (including initial 482 disturbance severity, perception, and competition) can vary through a stands recovery, 483 and thus effect the signals of recovery observed using Landsat time series. For example, 484 the apparent linearity of our recovery trajectories is due to the selection for recovering 485 stands that experience no further disturbances within our time series. In reality the 486 recovery process will vary based upon the nature and amount of vegetation present 487 over the course of forest recovery over time and could result in more complex and 488 subtlety varied recovery trajectories. The within pixel averaging of the vegetation 489 conditions will also serve to mute the year-on-year variability in spectral indices 490 491 relating the vegetation recovery. Likwise, forest recovery monitoring using passive optical remote sensing are, by definition, limited to observing changes in reflected light 492 493 from forest canopies. Previous efforts have demonstrated the difficulty inherent in 494 assessing successional state from remote sensing datasets (Song et al 2007, Schroeder et al 2006, Song et al 2003). As a result, the spectral forest recovery trajectories do not 495 necessarily directly correspond to changes in forest successional stages or may only be 496
- 497 able to report on a small amount of variation in biophysical variables. Rather spectral
- 498 trajectories essentially track the reestablishment of vegetation towards canopy closure
- 499 or when the spectral signals of a recovering forest stand resemble that of nearby

undisturbed vegetation or its previous state before the disturbance occurred (Kennedyet al 2012, Schroeder et al 2007, Buma 2012).

502

In this research we found that less than 1% of disturbed forests in both Boreal Shield 503 sections showed no signs of recovery, with nearly all disturbed forests showing signs of 504 recovery within 5 years. This implies that Boreal Shield forests overwhelmingly initiate 505 recovery processes after stand replacing disturbances. This is in agreement with 506 previous descriptions disturbance and subsequent recovery as key processes that shape 507 the patchwork of forest structures, types, and ages critical to maintaining the diversity 508 of the Canadian boreal zone (Bonan and Shugart 1989, Heinselman 1981). The spectral 509 recovery trajectories constructed in this paper depend upon changes in Wetness 510 through time. The shortwave infrared bands are weighted heavily in the calculation of 511 Wetness when compared to the visible and near infrared bands (Crist et al 1984). As a 512 result it is well suited to detecting changes from disturbance events such as fire. If a 513 514 disturbance or recovery event produces an observable change in other portions of the 515 spectrum other than the shortwave infrared, it is possible that this change may not be detected by the temporal segmentation algorithm. Thus care should be taken when 516 517 selecting a spectral index and knowledge of the disturbance types of interest and their characteristics is useful. It is important to recall this link between the spectral channels 518 used in a given index (or weight given in a combined index) with what is portrayed as 519 recovery. An index will likely cross a non-disturbed forest threshold sooner with a 520 visible and near-infrared driven index than an index driven by longer wavelengths. In 521 this research, we have been conservative by focusing on the SWIR-driven Wetness 522 index which has been shown to reflect changes in structural complexity and shadowing 523 and is therefore more sensitive to the presence of trees, than early successional herbs 524 and shrubs (Cohen et al 1995, Collins & Woodcock 1996, Franklin et al 2000). 525 Alternatively, initial increases in vegetated cover are well characterized with shorter 526 527 visible and near-infrared wavelengths (as used in Greenness). This more rapid saturation of the visible and near-infrared driven indices (i.e. Greenness) offers initial 528 insight that vegetation is present but has less capacity for providing information on 529 530 increasing structural complexity (Franklin et al 2000, Franklin et al 2002, Skakun et al 2003, Healey et al 2006). 531 532

532 533

### Case Study: Partitioning of the East and West Boreal Shield

We were able to characterize forest recovery using Landsat time series, focusing our
work on the known east/west division of the Boreal Shield ecozone. One of the
rationales for splitting the Boreal Shield ecozone into two distinct features was that the
regions experienced different ecoclimatic conditions, disturbance regimes, and forest
processes. Using our temporally segmented datasets describing the disturbance and
recovery trends in the both Boreal Shield sections, we suggest that characteristics of

- 540 disturbance and recovery for the east and west Boreal Shield study areas are also
- 541 different, further reinforcing the division of this ecozone.
- 542

543 The annual area disturbed shows different patterns of disturbance between the east and

- the west sections. From 1985-2000 there is a relatively constant area disturbed in the
- east. In contrast the west has a much more variable pattern of disturbance over the same
- time period (Figure 6). Interestingly, both the east and west show the same increasing
- 547 disturbance trends in the 2001-2012 periods, but at differing amounts with more
- 548 disturbances in the east than west.
- 549

550 The recovery trends of most eastern and western spectral indices follow the expected

- trend of an initial low value rising over time. However, Brightness trajectories are
- anticipated to begin with an initial high value and decrease over time. When the east
- and west Brightness section's trajectories are compared, large differences become
- apparent; the eastern trajectory follows a decreasing trend, while the western trajectory
- show no clear trend. We interpret this difference as related to the type and
- characteristics of disturbance that is most likely occurring. Bergeron et al (2004)
- 557demonstrated that fires are less frequent in the eastern than the western Boreal Shield
- and fire size is on average much larger in the west (Stocks et al 2002), which can lead to
- 559 more patchiness and mixed severity effects. The clear decreasing pattern of Brightness
- 560 in the east implies disturbances that are homogeneous entirely clearing a forest stand
- and leaving few if any remnant patches resulting in a uniform recovery signal. The lack
- of clear Brightness recovery trajectory in the west implies a more heterogeneoussituation with mixed fire severities leaving more remnant patches that diminish the
- 564 Brightness signal through the presence of more vegetation and decreased reflectivity in
- 565 the visible wavelengths.
- 566

567 Spectral recovery trajectory patterns relative to their stable mean also highlight differences in the east and west sections. For example most eastern trajectories increase 568 to their stable mean values (e.g. Wetness) and in some cases exceed their stable mean 569 570 (e.g. Greenness). In contrast the western trajectories rarely exceed their stable mean, with most trajectories remaining below the mean throughout the analysis period (e.g. 571 572 Wetness). We suggest these spectral recovery differences also relate known patterns of 573 forest stand recovery between the two sections. After a stand replacing disturbance, eastern Boreal Shield forest stands have been shown to initiate with regenerating broad 574

- 575 leaf shade intolerant treed vegetation, which have higher Wetness, and Greenness levels
- 576 than the shade tolerant conifers that often replace as the stand matures (Bergeron et al
- 577 2000, Bergeron et al 2001, Bergeron et al 2007, Chen and Popadiouk 2002). After a stand
- 578 replacing disturbance, western Boreal Shield stands typically initiate from the surviving
- 579 seedbank and mature towards even aged shade tolerant conifers or more complex and

580 layered conifer stands if another stand replacing disturbance does not occur (Bergeron

et al 2000, Brassard and Chen 2006). This pattern is shown in the western Boreal Shield

582 in the spectral recovery trajectory with lower Greenness, and Wetness values

- 583 corresponding to the coniferous dominated re-initiating stand.
- 584 585

The classifications of recovery rates show differences between the east and west (Figures 8 and 9). While both areas show an increase in the above average recovery rate towards the end of their time series, this is more evident in the east than the west. Also the amount of average recovery in the western study areas remains relatively stable from year 14, the midpoint of our analysis period, until the end of the time series, while the eastern Boreal Shield study areas average recovery increases over the same period of time.

593

The Cohen's *d* statistic enabled the direct comparison of recovery patterns between the 594 eastern and western Boreal Shield (Figures 3b-6b, Table 2). Three of four Cohen's *d* east 595 to west comparison trajectories commence within zones of small or moderate difference 596 and end with large difference between the two sections, indicating that their recovery is 597 observably different. The difference between the east and west in Wetness recovery is 598 599 especially important because it is sensitive to changes in forest structure. Noticeably, Wetness's comparison *d* trajectory shows the most change from its initial point of no 600 601 difference (*d* value 0.01) ending with large differences (*d* value 0.90). This essentially 602 means that although stand clearing disturbances may reduce the signal in disturbed areas in the east and west to similar Wetness levels, they recover from that level in 603 different manners, as indicated by their Wetness recovery trajectories and their Cohen's 604 605 d trajectories. Finally the comparison d trajectory shows that by the end of the time series, that east and west Boreal Shield recovery is largely (d>0.8) different. 606 It is important to recognize the factors that drive the variability in recovery between 607 these two sections of the Boreal Shield ecozone. A known precipitation gradient exists, 608 along with differences in winter average minimum temperatures, contributing to 609 differing climates (Urquizo et al 2000, Ecological Stratification Working Group 1996) 610 and as a result observed differences in vegetation recovery rates. Differences in 611 disturbance frequency and size is already well known (Stocks et al 2002) for the two 612 Boreal Shield sections which will also impact the observed spectral forest recovery. 613 Additionally, soil conditions could also affect forest recovery rates (Lecomte and 614 Bergeron 2005). The status and variability of all these factors combine to drive the 615 observed differences in the east and west spectral forest recovery trajectories. 616

617

### 619 Conclusion:

620 The objective of this study was to develop and apply a methodology for using Landsat

621 time series to characterize forest recovery using spectral recovery trajectories. To better

- understand spectral forest recovery, we selected and temporally segmented 1220
- 623 images from eight Landsat scenes to act as sample areas for the years 1984 through 2013
- and then extracted signals of post-disturbance recovery. We constructed, characterized,
- and then compared spectral forest recovery trajectories in four spectral indices between
- both east and west Boreal Shield ecozone sections and examined difference levels in
- recovery with the Cohen's *d* statistic. We also directly compared the recovery
- trajectories of four spectral indices derived from the post disturbed areas within the east
- and west sections to each other showing similarities and differences.

630

631 Using yearly Landsat time series we characterized spectral recovery and further

632 identified differences in recovery between the sections. We conclude that Landsat time

633 series are an effective dataset to characterize disturbance and the subsequent recovery

of forest vegetation. Spectral recovery is detectable, recovery trajectories are observable,

and characterizations following a stand replacing disturbance provide insight into

636 vegetative recovery and rates of recovery. Spatial trends of disturbance and recovery

are evident and differ highlighting the differences between the two Boreal Shield

sections. However, care must be taken when selecting a spectral index to temporally

639 segment a time series as some indices capture early vegetation return and often saturate

- 640 shortly thereafter. These indices therefore offer less information on post-disturbance
- 641 vegetation return and related increases in vegetation complexity. Also important is the

642 definition of recovery, as a process, and not necessarily a state. Further, the conditions

that are sampled and used to indicate recovery must be clearly stated as the ability toachieve or resemble these conditions is a function of the strata that was created. In this

645 particular research we created a strata of mature (stable, unchanged during the Landsat

646 TM, ETM+, satellite record) forest conditions to compare to the spectral conditions

647 present following stand replacing disturbances. All but 1% of the area disturbed had a

648 positive spectral trajectory following disturbance, implying that 99% of the forest area

649 disturbed is revegetating and initiating early successional processes leading towards a

- 650 treed state.
- 651
- 652

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- 659
- 660

#### References 661

662							
663	Andrew, M. E., Wulder, M. A., & Coops, N. C. (2012). Identification of de facto						
664	protected areas in boreal Canada. Biological Conservation, 146, 97–10						
665	Banskota, A., Kayastha, N., Falkowski, M. J., Wulder, M. A., Froese, R. E., & White, J. C.						
666	(2014). Forest monitoring using landsat time series data: a review. Canadian						
667	Journal of Remote Sensing, 40(5), 362-384.						
668	Baumann, M., Ozdogan, M., Kuemmerle, T., Wendland, K. J., Esipova, E., & Radeloff, V.						
669	C. (2012). Using the Landsat record to detect forest-cover changes during and after						
670	the collapse of the Soviet Union in the temperate zone of European Russia. Remote						
671	Sensing of Environment, 124, 174-184.						
672	Beck, P. S., & Goetz, S. J. (2011). Satellite observations of high northern latitude						
673	vegetation productivity changes between 1982 and 2008: ecological variability and						
674	regional differences. Environmental Research Letters, 6(4), 045501.						
675	Bergeron, Y. (2000). Species and stand dynamics in the mixed woods of Quebec's						
676	southern boreal forest. Ecology, 81(6), 1500-1516.						
677	Bergeron, Y., Drapeau, P., Gauthier, S., & Lecomte, N. (2007). Using knowledge of						
678	natural disturbances to support sustainable forest management in the northern						
679	Clay Belt. The Forestry Chronicle, 83(3), 326-337.						
680	Bergeron, Y., Flannigan, M., Gauthier, S., Leduc, A., & Lefort, P. (2004). Past, current						
681	and future fire frequency in the Canadian boreal forest: implications for						
682	sustainable forest management. AMBIO: A Journal of the Human						
683	Environment,33(6), 356-360.						
684	Bergeron, Y., Gauthier, S., Kafka, V., Lefort, P., & Lesieur, D. (2001). Natural fire						
685	frequency for the eastern Canadian boreal forest: consequences for sustainable						
686	forestry. Canadian Journal of Forest Research, 31(3), 384-391.						
687	Berner, L. T., Beck, P. S., Bunn, A. G., Lloyd, A. H., & Goetz, S. J. (2011). High-latitude						
688	tree growth and satellite vegetation indices: Correlations and trends in Russia and						
689	Canada (1982–2008). Journal of Geophysical Research: Biogeosciences (2005–						
690	2012), 116(G1).						

- Boisvenue, C., & Running, S. W. (2006). Impacts of climate change on natural forest
  productivity–evidence since the middle of the 20th century. Global Change
  Biology, 12(5), 862-882.
- Bolton, D. K., Coops, N. C., & Wulder, M. A. (2013). Measuring forest structure along
  productivity gradients in the Canadian boreal with small-footprint Lidar.
  Environmental Monitoring and Assessment, 185(8), 6617-6634.
- Bonan, G. B., & Shugart, H. H. (1989). Environmental factors and ecological processes in
  boreal forests. Annual Review of Ecology and Systematics, 1-28.
- Brandt, J. P., Flannigan, M. D., Maynard, D. G., Thompson, I. D., & Volney, W. J. A.
  (2013). An introduction to Canada's boreal zone: ecosystem processes, health,
  sustainability, and environmental issues 1. Environmental Reviews,21(4), 207-226
- 702 Brandt, J. P. (2009). The extent of the North American boreal zone. Environmental

703 Reviews, 17, 101-161.

- Brassard, B. W., & Chen, H. Y. (2006). Stand structural dynamics of North American
  boreal forests. Critical Reviews in Plant Sciences, 25(2), 115-137.
- Buma, B. (2012). Evaluating the utility and seasonality of NDVI values for assessing
  post-disturbance recovery in a subalpine forest. Environmental Monitoring and
  Assessment, 184(6), 3849-3860.
- Cavin, L., Mountford, E. P., Peterken, G. F., & Jump, A. S. (2013). Extreme drought alters
  competitive dominance within and between tree species in a mixed forest
  stand. Functional Ecology, 27(6), 1424-1435.
- Chen, H. Y., & Popadiouk, R. V. (2002). Dynamics of North American boreal
  mixedwoods. Environmental Reviews, 10(3), 137-166.
- Chen, X., Vogelmann, J. E., Rollins, M., Ohlen, D., Key, C. H., Yang, L., ... & Shi, H.
- 715 (2011). Detecting post-fire burn severity and vegetation recovery using
- 716 multitemporal remote sensing spectral indices and field-collected composite burn
- 717 index data in a ponderosa pine forest. International Journal of Remote
- **718**Sensing, 32(23), 7905-7927.
- Chu, T., & Guo, X. (2013). Remote sensing techniques in monitoring post-fire effects and
  patterns of forest recovery in boreal forest regions: a review. Remote Sensing, 6(1),
  470-520.
- Cohen, J. (1988) Statistical power analysis for the behavioral sciences. 2nd Ed. Hove:Lawrence Erlbaum Associates.
- Cohen, W. B., & Spies, T. A. (1992). Estimating structural attributes of Douglasfir/western hemlock forest stands from Landsat and SPOT imagery. Remote
  Sensing of Environment, 41(1), 1-17.
- Cohen, W. B., Fiorella, M., Gray, J., Helmer, E., & Anderson, K. (1998). An efficient and
   accurate method for mapping forest clearcuts in the Pacific Northwest using
- 729Landsat imagery. Photogrammetric Engineering and Remote Sensing, 64(4), 293-

- Cohen, W. B., Spies, T. A., & Fiorella, M. (1995). Estimating the age and structure of
  forests in a multi-ownership landscape of western Oregon, USA. International
  Journal of Remote Sensing, 16(4), 721-746.
- Collins, J. B., & Woodcock, C. E. (1996). An assessment of several linear change
  detection techniques for mapping forest mortality using multitemporal Landsat
  TM data. Remote Sensing of Environment, 56(1), 66–77.
- 737 Crist, E. P., & Cicone, R. C. (1984). A physically-based transformation of Thematic
  738 Mapper data---The TM Tasseled Cap. IEEE Transactions on Geoscience and
- 739 Remote Sensing, IEEE Transactions On, (3), 256-263.
  740 Cuevas-Gonzalez, M., Gerard, F., Balzter, H., & Riano, D. (2009). Analysing forest
- recovery after wildfire disturbance in boreal Siberia using remotely sensed
  vegetation indices. Global Change Biology, 15(3), 561-577.
- 743 Ecological Stratification Working Group. (1996). A national ecological framework for
- 744 Canada. Agriculture and Agri-Food Canada and Environment Canada: Ottawa,745 ON. Available from:
- http://sis.agr.gc.ca/cansis/publications/ecostrat/cad\_report.pdf [cited on May 12<sup>th</sup>
  2015]
- Fidenshink, J., Schwind, B., Brewer, K., Zhu, Z., Quayle, B., & Howard, S. (2007). A
  project for monitoring trends in burn severity. Fire Ecology 3 (1): 3-21.Fire Ecology
  Special Issue Vol, 3, 4.
- For Epting, J., Verbyla, D., & Sorbel, B. (2005). Evaluation of remotely sensed indices for
  assessing burn severity in interior Alaska using Landsat TM and ETM+. Remote
  Sensing of Environment, 96(3), 328-339.
- Flannigan, M. D., Logan, K. A., Amiro, B. D., Skinner, W. R., & Stocks, B. J. (2005).
  Future area burned in Canada. Climatic Change, 72(1-2), 1-16
- Franklin, S. E., Moskal, L. M., Lavigne, M. B., & Pugh, K. (2000). Interpretation and
  classification of partially harvested forest stands in the Fundy Model Forest using
  multitemporal Landsat TM. Canadian Journal of Remote Sensing, 26, 318–333.
- Franklin, S.E., Lavigne, M.B., Wulder, M.A., & Stenhouse, G.B. (2002). Change detection
  and landscape structure mapping using remote sensing. The Forestry Chronicle.
  761 78(5), 618-625
- Frazier, R. J., Coops, N. C., Wulder, M. A., & Kennedy, R. (2014). Characterization of
  aboveground biomass in an unmanaged boreal forest using Landsat temporal
  segmentation metrics. ISPRS Journal of Photogrammetry and Remote Sensing, 92,
- 765 137-146.
- 766 Frolking, S., Palace, M. W., Clark, D. B., Chambers, J. Q., Shugart, H. H., & Hurtt, G. C.
- 767 (2009). Forest disturbance and recovery: A general review in the context of
- spaceborne remote sensing of impacts on aboveground biomass and canopy
- structure. Journal of Geophysical Research: Biogeosciences (2005–2012), 114(G2).

- Gitas, I., Polychronaki, A., Mitri, G., & Veraverbeke, S. (2012). Advances in remote
  sensing of post-fire vegetation recovery monitoring-a review. INTECH Open
  Access Publisher.
- Goetz, S. J., Fiske, G. J., & Bunn, A. G. (2006). Using satellite time-series data sets to
  analyze fire disturbance and forest recovery across Canada. Remote Sensing of
  Environment, 101(3), 352-365.
- Griffiths, P., Kuemmerle, T., Baumann, M., Radeloff, V. C., Abrudan, I. V., Lieskovsky,
  J., ... & Hostert, P. (2014). Forest disturbances, forest recovery, and changes in
  forest types across the Carpathian ecoregion from 1985 to 2010 based on Landsat
  image composites. Remote Sensing of Environment, 151, 72-88.
- Guindon, L., Bernier, P. Y., Beaudoin, A., Pouliot, D., Villemaire, P., Hall, R. J., ... & StAmant, R. (2014). Annual mapping of large forest disturbances across Canada's
  forests using 250 m MODIS imagery from 2000 to 2011. Canadian Journal of Forest

783 Research, 44(12), 1545-1554.

- Hansen, M. C., Potapov, P. V., Moore, R., Hancher, M., Turubanova, S. A., Tyukavina,
  A., ... & Townshend, J. R. G. (2013). High-resolution global maps of 21st-century
  forest cover change. Science, 342(6160), 850-853.
- Hansen, M. J., Franklin, S. E., Woudsma, C., & Peterson, M. (2001). Forest structure
  classification in the North Columbia mountains using the Landsat TM Tasseled
  Cap Wetness component. Canadian Journal of Remote Sensing, 27(1), 20-32.
- Healey, S. P., Cohen, W. B., Zhiqiang, Y., & Krankina, O. N. (2005). Comparison of
  Tasseled Cap-based Landsat data structures for use in forest disturbance
  detection. Remote Sensing of Environment, 97(3), 301-310.
- Healey, S. P., Yang, Z., Cohen, W. B., & Pierce, D. J. (2006). Application of two
  regression-based methods to estimate the effects of partial harvest on forest
- structure using Landsat data. Remote Sensing of Environment, 101(1), 115-126.
- Heinselman, M. L. (1981). Fire and succession in the conifer forests of northern North
  America. In Forest succession (pp. 374-405). Springer New York.
- Hermosilla, T., Wulder, M. A., White, J. C., Coops, N. C., & Hobart, G. W. (2015). An
  integrated Landsat time series protocol for change detection and generation of
  annual gap-free surface reflectance composites. Remote Sensing of

801 Environment, 158, 220-234

- Huang, C., Goward, S. N., Masek, J. G., Thomas, N., Zhu, Z., & Vogelmann, J. E. (2010).
  An automated approach for reconstructing recent forest disturbance history using
  dense Landsat time series stacks. Remote Sensing of Environment, 114(1), 183-198.
- Huang, C., Wylie, B., Yang, L., Homer, C., & Zylstra, G. (2002). Derivation of a Tasselled
- 806 Cap transformation based on Landsat 7 at-satellite reflectance. International
  807 Journal of Remote Sensing, 23(8), 1741-1748.

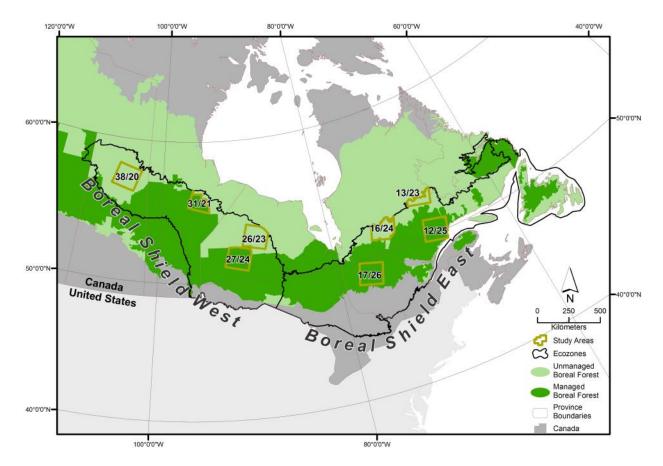
- Huete, A. R., Liu, H. Q., Batchily, K., & Van Leeuwen, W. J. D. A. (1997). A comparison
  of vegetation indices over a global set of TM images for EOS-MODIS. Remote
  Sensing of Environment, 59(3), 440-451.
- Johnson, D. H. (1999). The insignificance of statistical significance testing. The Journal of
  Wildlife Management, 763-772.
- Johnson, E. A. (1996). Fire and vegetation dynamics: studies from the North Americanboreal forest. Cambridge University Press.
- Johnstone, J. F., & Chapin III, F. S. (2006a). Effects of soil burn severity on post-fire tree
  recruitment in boreal forest. Ecosystems, 9(1), 14-31.
- Johnstone, J. F., & Chapin III, F. S. (2006b). Fire interval effects on successional trajectory
  in boreal forests of northwest Canada. Ecosystems, 9(2), 268-277.
- 819 Kasischke, E. S., & Turetsky, M. R. (2006). Recent changes in the fire regime across the
- North American boreal region—spatial and temporal patterns of burning across
  Canada and Alaska. Geophysical Research Letters, 33(9).
- Kasischke, E. S., Loboda, T., Giglio, L., French, N. H., Hoy, E. E., de Jong, B., & Riano, D.
- (2011). Quantifying burned area for North American forests: Implications for
  direct reduction of carbon stocks. Journal of Geophysical Research: Biogeosciences
  (2005–2012), 116(G4).
- Kennedy, R. E., Yang, Z., & Cohen, W. B. (2010). Detecting trends in forest disturbance
  and recovery using yearly Landsat time series: 1. LandTrendr Temporal
  segmentation algorithms. Remote Sensing of Environment, 114(12), 2897-2910.
- Kennedy, R. E., Yang, Z., Cohen, W. B., Pfaff, E., Braaten, J., & Nelson, P. (2012). Spatial
  and temporal patterns of forest disturbance and regrowth within the area of the
  Nertherest Exact Plan. Provets Consists of Environment, 122, 117, 122.
- 831 Northwest Forest Plan. Remote Sensing of Environment, 122, 117-133.
- Kull, S. J., Kurz, W. A., Rampley, G. J., Banfield, G. E., Schivatcheva, R. K., & Apps, M. J.
  (2006). Operational-scale carbon budget model of the Canadian forest sector
  (CBM-CFS3) version 1.0: user's guide. 2006.
- Kurz, W.A.; Apps, M.J.; Webb, T.M.; McNamee, P.J. (1992). The carbon budget of the
  Canadian forest sector: Phase I. For. Can., Northwest Reg., Edmonton, AB. Inf.
  Rep. NOR-X-326
- LeComte, N., Bergeron, Y. (2005). Successional pathways on different surficial deposits
  in the coniferous boreal forest of the Quebec Clay Belt. Canadian Journal of Forest
  Research, 2005, vol. 35, no 8, p. 1984-1995.
- Lentile, L. B., Holden, Z. A., Smith, A. M., Falkowski, M. J., Hudak, A. T., Morgan, P., ...
  & Benson, N. C. (2006). Remote sensing techniques to assess active fire
  characteristics and post-fire effects. International Journal of Wildland Fire, 15(3),
  319-345.
- Loboda, T. V., Zhang, Z., O'Neal, K. J., Sun, G., Csiszar, I. A., Shugart, H. H., &
  Sherman, N. J. (2012). Reconstructing disturbance history using satellite-based

assessment of the distribution of land cover in the Russian Far East. Remote 847 Sensing of Environment, 118, 241-248. 848 Masek, J. G. (2001). Stability of boreal forest stands during recent climate change: 849 evidence from Landsat satellite imagery. Journal of Biogeography, 28(8), 967-976. 850 Masek, J. G., Goward, S. N., Kennedy, R. E., Cohen, W. B., Moisen, G. G., Schleeweis, K., 851 & Huang, C. (2013). United States forest disturbance trends observed using 852 Landsat time series. Ecosystems, 16(6), 1087-1104. 853 Masek, J.G., E.F. Vermote, N. Saleous, R. Wolfe, F.G. Hall, Huemmrich, K.F., Gao, J. 854 855 Kutler, and T.K. Lim, (2006), A Landsat surface reflectance data set for North America, 1990-2000, Geoscience and Remote Sensing Letters, 3, 68-72. 856 Mehrabi, Z., Slade, E. M., Solis, A., & Mann, D. J. (2014). The importance of microhabitat 857 for biodiversity sampling. PloS one, 9(12), e114015. 858 Nakagawa, S., & Cuthill, I. C. (2007). Effect size, confidence interval and statistical 859 significance: a practical guide for biologists. Biological Reviews, 82(4), 591-605. 860 861 Natural Resources Canada. (2014). Boreal forest. Available from: https://www.nrcan.gc.ca/forests/boreal/13071. [Cited on May 12 2015]. 862 Oliver, C. D., & Larson, B. C. (1990). Forest stand dynamics. McGraw-Hill, Inc.. 863 864 Payette, S. (1992). Fire as a controlling process in the North American boreal forest. A systems analysis of the global boreal forest, 144-169. 865 Pflugmacher, D., Cohen, W.B., & Kennedy, R.E. (2012). Using Landsat derived 866 disturbance history (1972-2012) to predict current forest structure. Remote Sensing 867 of Environment, 122, 146-165. 868 Pickell, P.D., Hermosilla, T, Frazier, R.J., Coops, N.C., Wulder, M.A. (2015). Forest 869 recovery trends derived from Landsat time series for North American boreal 870 871 forests. IN PREP. Potapov, P. V., Turubanova, S. A., Tyukavina, A., Krylov, A. M., McCarty, J. L., 872 Radeloff, V. C., & Hansen, M. C. (2014). Eastern Europe's forest cover dynamics 873 874 from 1985 to 2012 quantified from the full Landsat archive. Remote Sensing of Environment. 875 Potapov, P., Hansen, M. C., Stehman, S. V., Loveland, T. R., & Pittman, K. (2008). 876 877 Combining MODIS and Landsat imagery to estimate and map boreal forest cover loss. Remote Sensing of Environment, 112(9), 3708-3719. 878 879 Powell, S. L., Cohen, W. B., Healey, S. P., Kennedy, R. E., Moisen, G. G., Pierce, K. B., & 880 Ohmann, J. L. (2010). Quantification of live aboveground forest biomass dynamics with Landsat time series and field inventory data: A comparison of empirical 881 modeling approaches. Remote Sensing of Environment, 114, 1053–1068. 882 883 Powers, R. P., Hermosilla, T., Coops, N. C., & Chen, G. (2015). Remote sensing and object-based techniques for mapping fine-scale industrial disturbances. 884 International Journal of Applied Earth Observation and Geoinformation, 34, 51-57. 885

- Price, D. T., Alfaro, R. I., Brown, K. J., Flannigan, M. D., Fleming, R. A., Hogg, E. H., ... &
  Venier, L. A. (2013). Anticipating the consequences of climate change for Canada's
  boreal forest ecosystems 1. Environmental Reviews,21(4), 322-365.
- Sader, S. A., Waide, R. B., Lawrence, W. T., & Joyce, A. T. (1989). Tropical forest biomass
  and successional age class relationships to a vegetation index derived from
- Landsat TM data. Remote Sensing of Environment, 28, 143-198.
- Schroeder, T. A., Cohen, W. B., & Yang, Z. (2007). Patterns of forest regrowth following
  clearcutting in western Oregon as determined from a Landsat time-series. Forest
  Ecology and Management, 243(2), 259-273.
- Schroeder, T. A., Cohen, W. B., Song, C., Canty, M. J., & Yang, Z. (2006). Radiometric
  correction of multi-temporal Landsat data for characterization of early
  successional forest patterns in western Oregon. Remote Sensing of
- 898 Environment, 103(1), 16-26.
- Schroeder, T. A., Wulder, M. A., Healey, S. P., & Moisen, G. G. (2011). Mapping wildfire
  and clearcut harvest disturbances in boreal forests with Landsat time series
  data. Remote Sensing of Environment, 115(6), 1421-1433.
- Shatford, J. P. A., Hibbs, D. E., & Puettmann, K. J. (2007). Conifer regeneration after
  forest fire in the Klamath-Siskiyous: How much, how soon?. Journal of
  Forestry, 105(3), 139-146.
- Shvidenko, A., & Apps, M. (2006). The International Boreal Forest Research Association:
  understanding boreal forests and forestry in a changing world. Mitigation and
  Adaptation Strategies for Global Change, 11(1), 5-32.
- Skakun, R. S., Wulder, M. A., & Franklin, S. E. (2003). Sensitivity of the Thematic
  Mapper enhanced wetness difference index to detect mountain pine beetle redattack damage. Remote Sensing of Environment, 86(4), 433–443.
- Song, C., & Woodcock, C. E. (2003). Monitoring forest succession with multitemporal
  Landsat images: factors of uncertainty. IEEE Transactions on Geoscience and
  Remote Sensing, 41(11), 2557-2567.
- Song, C., Schroeder, T. A., & Cohen, W. B. (2007). Predicting temperate conifer forest
  successional stage distributions with multitemporal Landsat Thematic Mapper
  imagery. Remote Sensing of Environment, 106(2), 228-237.
- 917 Steyaert, L. T., Hall, F. G., & Loveland, T. R. (1997). Land cover mapping, fire
- regeneration, and scaling studies in the Canadian boreal forest with 1 km AVHRR
  and Landsat TM data. Journal of Geophysical Research: Atmospheres (1984–
  2012), 102(D24), 29581-29598.
- Stocks, B. J., Mason, J. A., Todd, J. B., Bosch, E. M., Wotton, B. M., Amiro, B. D., ... &
  Skinner, W. R. (2002). Large forest fires in Canada, 1959–1997. Journal of
- 923 Geophysical Research: Atmospheres (1984–2012), 107(D1), FFR-5.
- 924 Urquizo, N., Bastedo, J., Brydges, T., & Shear, H. (2000). Ecological assessment of the
  925 boreal shield ecozone. Environment Canada, Ottawa, Ont. ISBN-662-28679-0.

Wesner, J. S., Billman, E. J., & Belk, M. C. (2012). Multiple predators indirectly alter 926 community assembly across ecological boundaries. Ecology, 93(7), 1674-1682. 927 Wiken, E.B. (1986). Terrestrial ecozones of Canada. Ecological Land Classification 928 Series No. 19. Environment Canada, Hull, Quebec. 929 Woodcock, C. E., Allen, R., Anderson, M., Belward, A., Bindschadler, R., Cohen, W., ... 930 & Wynne, R. (2008). Free access to Landsat imagery. Science (New York, 931 NY), 320(5879), 1011. 932 Wulder, M. A., Masek, J. G., Cohen, W. B., Loveland, T. R., & Woodcock, C. E. (2012). 933 934 Opening the archive: How free data has enabled the science and monitoring promise of Landsat. Remote Sensing of Environment, 122, 2-10. 935 Wulder, M. A., Ortlepp, S. M., White, J. C., & Maxwell, S. (2008). Evaluation of Landsat-936 7 SLC-off image products for forest change detection. Canadian Journal of Remote 937 Sensing, 34(2), 93-99. 938 939 Wulder, M. A., White, J. C., Alvarez, F., Han, T., Rogan, J., & Hawkes, B. (2009). 940 Characterizing boreal forest wildfire with multi-temporal Landsat and LIDAR data. Remote Sensing of Environment, 113(7), 1540-1555. 941 Wulder, M. A., & Seemann, D. (2001). Spatially partitioning Canada with the Landsat 942 943 worldwide referencing system. Canadian Journal of Remote Sensing, 27(3), 225-231. 944 Wulder, M.A., Campbell, C., White, J.C., Flannigan, M., & Campbell, I.D. (2007). 945 National circumstances in the international circumboreal community. The Forestry 946 Chronicle. 83(4), 539-556. 947 Xu, L., Myneni, R. B., Chapin Iii, F. S., Callaghan, T. V., Pinzon, J. E., Tucker, C. J., ... & 948 Stroeve, J. C. (2013). Temperature and vegetation seasonality diminishment over 949 950 northern lands. Nature Climate Change, 3(6), 581-586. Zhang, Q., Pavlic, G., Chen, W., Latifovic, R., Fraser, R., & Cihlar, J. (2004). Deriving 951 stand age distribution in boreal forests using SPOT Vegetation and NOAA 952 953 AVHRR imagery. Remote Sensing of Environment, 91(3), 405-418. Zhu, Z., & Woodcock, C. E. (2012). Object-based cloud and cloud shadow detection in 954 Landsat imagery. Remote Sensing of Environment, 118, 83-94. 955 956 Zhu, Z., Woodcock, C. E., & Olofsson, P. (2012). Continuous monitoring of forest 957 disturbance using all available Landsat imagery. Remote Sensing of Environment, 122, 75-91. 958 959

# 960 Figures



961

- 962 Figure 1: Boreal Shield East and West ecozones and forest management zones in
- 963 Canada with the eight study scene portions shown.

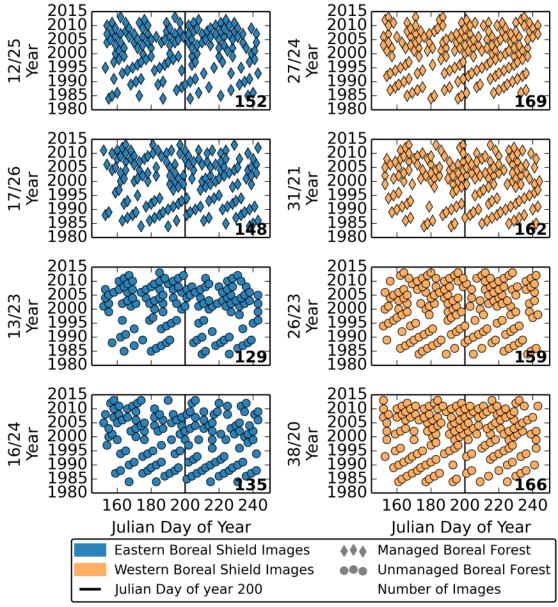
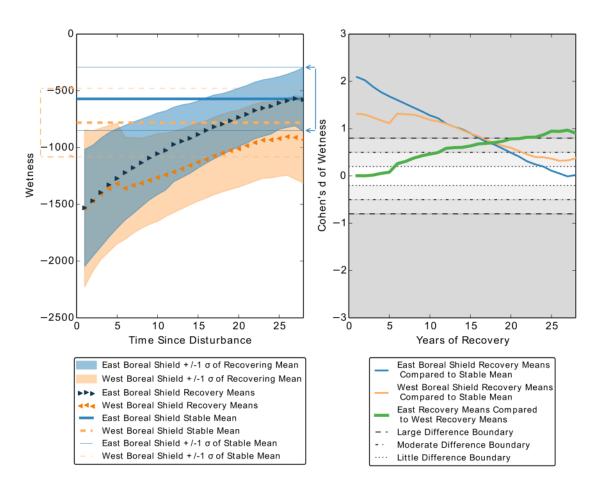
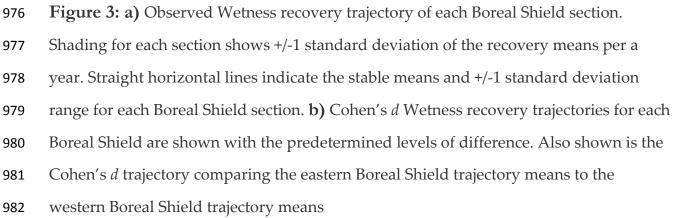




Figure 2: Distribution of dates and years of imagery for each Landsat footprint portion
used in this study. Eastern Boreal Shield path row scene footprints are shown in blue
while western Boreal Shield path row scene footprints are shown in orange. Diamond
shaped symbols indicate images located in the management zone, while circle shaped
symbols indicate images within the non-management zone. A total of 1220 images were
used divided between the 8 scene footprints.





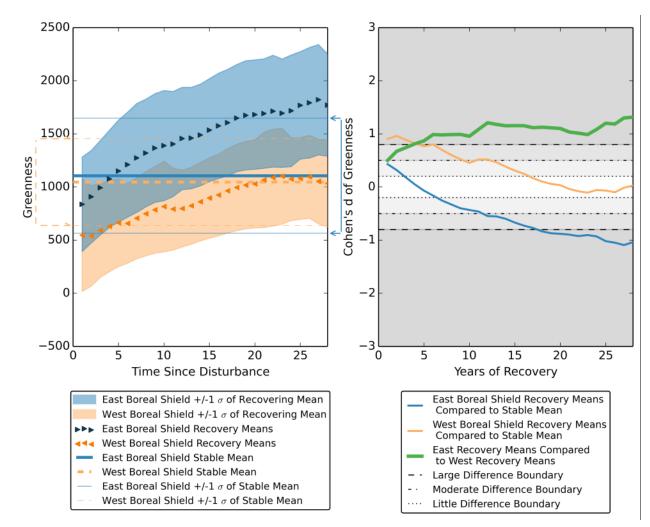


Figure 4: a) Observed Greenness recovery trajectory of each Boreal Shield section.
Shading for each section shows +/-1 standard deviation of the recovery means per a
year. Straight horizontal lines indicate the stable means and +/-1 standard deviation
range for each Boreal Shield section. b) Cohen's *d* Greenness recovery trajectories for
each Boreal Shield are shown with the predetermined levels of difference. Also shown
is the Cohen's *d* trajectory comparing the eastern Boreal Shield trajectory means to the
western Boreal Shield trajectory means

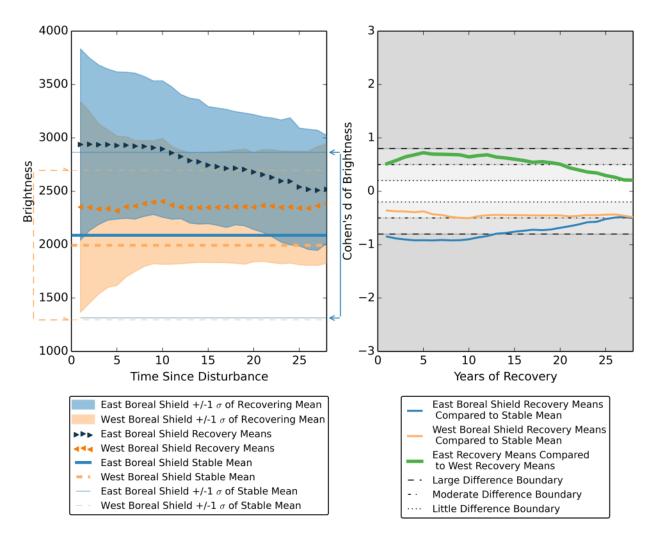


Figure 5: a) Observed Brightness recovery trajectory of each Boreal Shield section.
Shading for each section shows +/-1 standard deviation of the recovery means per a
year. Straight horizontal lines indicate the stable means and +/-1 standard deviation
range for each Boreal Shield section. b) Cohen's *d* Brightness recovery trajectories for
each Boreal Shield are shown with the predetermined levels of difference. Also shown
is the Cohen's *d* trajectory comparing the eastern Boreal Shield trajectory means to the
western Boreal Shield trajectory means.

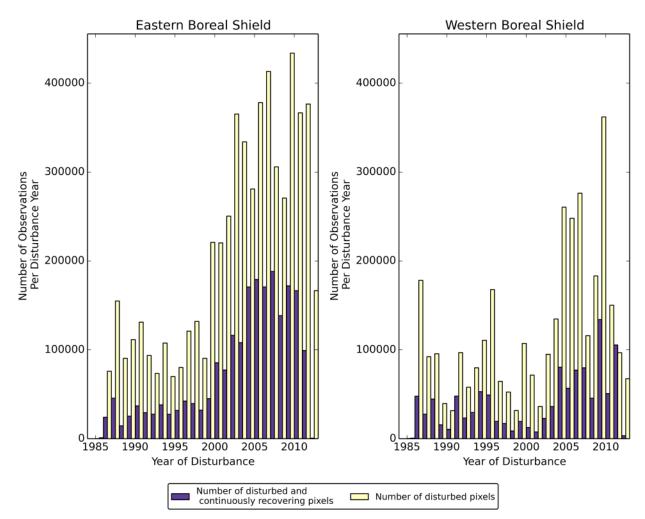


Figure 6: Number of disturbed pixels per year, and number of analysed pixels that
were disturbed and recover unabatedly through the end of their time series for the a)

1007 eastern study areas and **b**) western study areas.

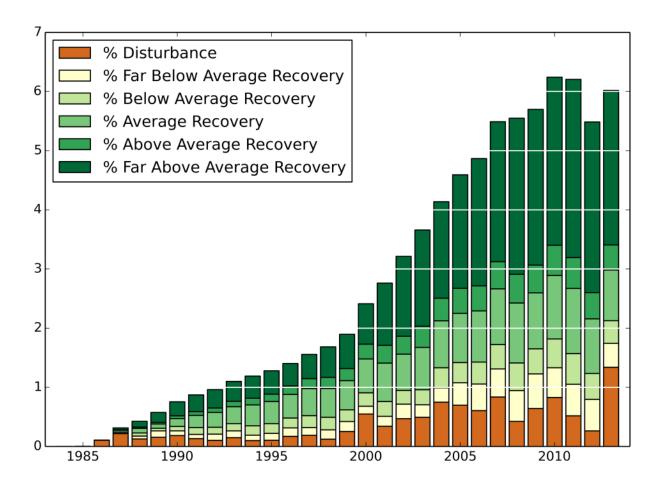


Figure 7: Boreal Shield East study areas spatial disturbance and recovery trends shown
as a percent of the total forest area in the study areas. Disturbance is calculated
annually. Recovery is cumulative and the categories of recovery are classified into five
categories based on the recovery trajectory means and standard deviations from the
Wetness trajectories in Figure 3a.

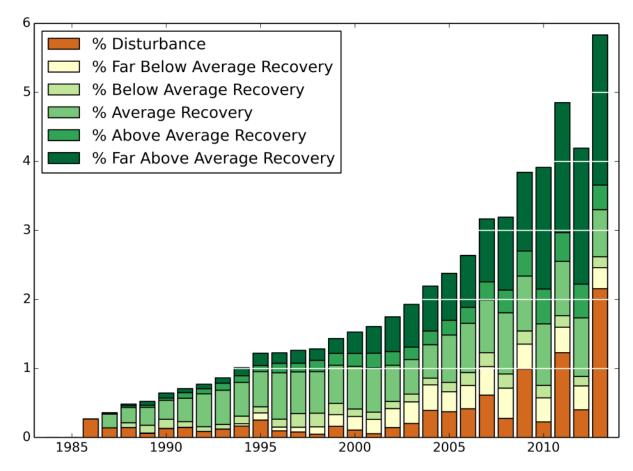




Figure 8: Boreal Shield West study areas spatial disturbance and recovery trends
shown as a percent of the total forest area in the study areas. Disturbance is calculated
annually. Recovery is cumulative and the categories of recovery are classified into five
categories based on the recovery trajectory means and standard deviations from the
Wetness trajectories in Figure 3a.

#### 1027 Tables

1028 **Table 1:** Summary statistics about each Landsat scene's management and inventory 1029 area, showing the individual path rows from which scene portions were used to 1030 construct the Landsat time series and the amount of boreal forest present in each scene. 1031 The amount of area experiencing each type of management paradigm in each Boreal Shield section and as a percent of the total forested area in that Boreal Shield section 1032 and management zone is also displayed in Table 1. Table 1 additionally enumerates the 1033 total area under study per a Boreal Shield section and as a percent of the total forested 1034 area within that Boreal Shield section. The values reflected in Table 1 show that we have 1035 1036 divided our study areas as evenly as possible between zones of management and Boreal Shield East and West sections. 1037

1038

Ecozone Name	Boreal Forest Managed/ Unmanaged	WRS2 Path and Row Portion	Boreal Forest Area in Scene (km²)	Study Zone Area per Management (km <sup>2</sup> )	Study Zone Area as a % of Ecozone per Management	Total Study Zone Area in Ecozone (KM^2)	Tota <b>l \$1349</b> Zone Area as a % of Ec <b>12046</b> 0	
Boreal	Managed	17/26 12/25	31649 30943	- 62592	7.60%	97865		1041
Shield East	Unmanaged	13/23	<u> </u>	35272	4.29%		<sup>11</sup> 1042	
		16/24	23701				1043	
	Managed	31/21	15742	46078	5.88%	105485	1045	
Boreal		27/24	30336				1347%4	
Shield West	Unmanaged	26/23	29305	59406	7.58%	105405	1045	
		38/20	30101				1045	
							1046	

1048
1049 Table 2: Summarization of Cohen's *d* trajectories by difference category and values at
1050 the start, midpoint, and end of the trajectory for Brightness, Greenness, and Wetness
1051 comparing east and west Boreal Shield recovery means to their stable means, and
1052 comparing eastern Boreal Shield recovery means to western Boreal Shield recovery
1053 means.

		East Recovery Means vs East No Change Stable Forest Mean			West Recovery Means vs West No Change Stable Forest Mean			East to West Comparison of Recovery Means		
	Recovery Year	1	14	28	1	14	28	1	14	28
	Brightness	-0.84	-0.74	-0.49	036	-0.45	-0.48	0.51	0.62	0.21
	Greenness	0.49	-0.60	-104	0.90	0.39	0.03	0.50	1.51	1.31
	Wetness	2.09	0.98	0.02	1.31	0.97	0.37	0.01	0.61	0.90
1055										_
	Levels o	f Difference: Large			Moderate Little		None			
1056										_
1057										